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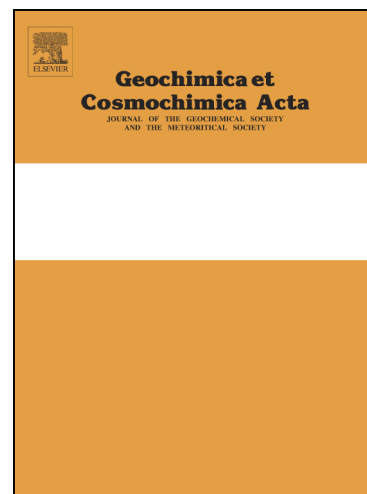
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# Geochemistry of dissolved aluminum at low pH: Extent and significance of Al-Fe<sup>(III)</sup> coprecipitation below pH 4.0

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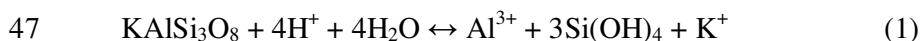
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**Abstract** – This work examines the geochemical behaviour of dissolved aluminum in sulfate-rich acidic waters. Our observations were obtained during several years of geochemical and mineralogical research in the San Telmo acidic pit lake and other pit lakes of SW Spain. The work includes scanning and transmission electron microscopy (SEM, TEM) of suspended mineral colloids found in deep lake waters. Energy dispersion spectroscopy (EDS) coupled to scanning and high resolution transmission electron microscopy (STEM, HRTEM) revealed not only the presence and formation of discrete, sub-micron Al solids like alunite, but also the abundance and distribution of Al into Fe(III) phases typical of acid mine drainage, such as schwertmannite and jarosite, at a nanometric resolution. The main conclusion emerging from our work is that the fate and transport of Al at low pH (<4.0) can be largely influenced by adsorption on and/or coprecipitation with both schwertmannite and jarosite. Under the geochemical conditions studied ( $\text{SO}_4^{2-}=10^{-2}$  M,  $\text{Fe(III)}\sim\text{Al}=10^{-3}$  M), alunite formation may occur at  $\text{pH}>3.3$ , as suggested by mineralogical observations and geochemical modelling. Below this pH, and contrary to the extended assumption, Al is not truly conservative, and in the presence of ferric iron, both metals may co-precipitate at a substantial extent to form either particles of Al-rich schwertmannite (containing up to ca. 8 at.% Al with  $[\text{Fe}/(\text{Fe}+\text{Al})]=0.77$ ) and/or crystals of  $\text{H}_3\text{O}^+$ - to  $\text{K}^+$ -jarosite (containing up to ca. 10 at.% Al with  $[\text{Fe}/(\text{Fe}+\text{Al})]=0.54$ ). This Al incorporation seems to take place by adsorption on particle surfaces in schwertmannite and by atomic substitution for  $\text{Fe}^{3+}$  in jarosite. Alunite is also unstable at this low pH range with respect to jarosite, which may lead either to isomorphic transformation and/or to chemically zoned crystals with jarositic rims around previously formed alunite cores. As a whole, the compositional spectrum of the analysed jarosites and alunites describes a discontinuous, coupled ( $\text{Al}^{3+}\text{-Fe}^{3+}$ ,  $\text{H}_3\text{O}^+\text{-K}^+$ ) solid solution series with an apparent gap at intermediate compositions. However, this gap seems to follow geochemical aspects more than crystallographic factors (i.e., immiscibility). The combination of this macroscopically *invisible* Al incorporated into Fe(III) solids along with subordinate alunite formation may cause significant Al removal even at very low pH (e.g., 20% decrease in Al concentration in San Telmo at  $\text{pH}<3.1$ ). Furthermore, this Fe(III)-Al co-precipitation may also affect the fate of toxic trace elements like As and Pb.

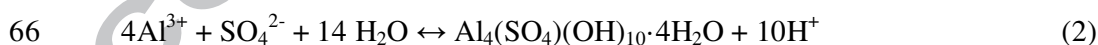
44 **1. INTRODUCTION**

45 Aluminum is present in acidic waters around the world. The main source of aluminum in  
46 these environments is the dissolution of aluminosilicates (*e.g.*, feldspar; reaction 1):



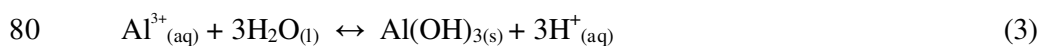
48 The release of Al may be enhanced during microbially-mediated pyrite oxidation by the  
49 generation of sulfuric acid (Nordstrom and Alpers, 1999; Bigham and Nordstrom, 2000).  
50 Given its high toxicity to many organisms (*e.g.*, Garciduenas and Cervantes, 1996), the  
51 geochemical behaviour of aluminum in aquatic systems has been widely studied. In particular,  
52 the geochemistry of aluminum in acidic waters (*e.g.*, mine waters) and soils has been  
53 addressed in numerous studies (*e.g.*, Van Bremen et al., 1973; Nordstrom, 1982; Nordstrom  
54 and Ball, 1986; Karathanasis et al., 1988; Lindsay and Walthall, 1996; Nordstrom and Alpers,  
55 1999; Bigham and Nordstrom, 2000; Sánchez-España, 2007; Jones et al., 2011). Dissolved  
56 aluminum in fresh waters is mostly present in the form of Al-OH species (*e.g.*,  $\text{Al}(\text{OH})^{2+}$ ,  
57  $\text{Al}(\text{OH})_2^+$ ). Below  $\text{pH} < 5.0$ , aluminum is usually considered to remain in solution and, above  
58 this pH, to precipitate as Al hydroxides such as gibbsite or amorphous  $\text{Al}(\text{OH})_3$ . In contrast,  
59 aluminum speciation in acid mine drainage (AMD) is strongly influenced by the high  
60 concentration of sulfate ( $\text{SO}_4^{2-}$ ) such that sulfate complexes (*i.e.*,  $\text{AlSO}_4^+$ ,  $\text{Al}(\text{SO}_4)^{2-}$ ) are  
61 predominant compared to  $\text{Al}^{3+}$ .

62 Aluminum buffers many acidic (pit) lakes, where experimental and modelling work has  
63 shown that  $\text{Al}^{3+}$  may precipitate at  $\text{pH} > 4.0$  near the lake bottom (Sánchez-España et al., 2011).  
64 At this pH value, nanocrystalline oxyhydroxysulfate minerals such as felsöbányaite (or its  
65 more hydrated precursor, hydrobasaluminite) may form by reaction 2:



67 The formation of these colloidal solids has been widely observed in acidic surface waters  
68 of the Iberian Pyrite Belt (IPB) mining district, where they promote significant removal of  
69 dissolved Al (Sánchez-España et al., 2005, 2006). Other papers have addressed the role of  
70 alunite ( $\text{KAl}_3(\text{SO}_4)_2(\text{OH})_6$ ) on the control of Al concentration in acidified surface waters and  
71 soils (*e.g.*, Adams and Rawajfih, 1977; Nordstrom, 1982; Singh, 1982), and geochemical  
72 modelling calculations usually suggest saturation of many acid pit lakes with respect to this  
73 mineral (Eary, 1999; Sánchez-España et al., 2011; Eary and Castendyk, 2013).

74 A number of likely controls on Al solubility have been proposed in surface waters. Gibbsite  
75 and kaolinite solubilities usually control aluminum concentration in natural waters, and the  
76 behaviour of Al in acidic waters has traditionally been explained by the precipitation of  
77 microcrystalline gibbsite or amorphous  $\text{Al}(\text{OH})_3$  through reaction 3 (Nordstrom, 1982;  
78 Driscoll et al., 1984; Nordstrom and Ball, 1986; Hendershot et al., 1996; Nordstrom and  
79 Alpers, 1999):



81 The presence of dissolved sulfate modifies the behaviour of aluminum under acidic  
82 conditions, and the solubility of this metal is controlled by some other minerals (such as  
83 alunite or felsobanyaite) which can precipitate at lower pH with respect to near-neutral waters  
84 with little or no dissolved sulfate (Nordstrom, 1982; Bertsch and Parker, 1996; Bigham and  
85 Nordstrom, 2000). Thus, the transition between conservative (i.e., dissolved concentrations  
86 remaining constant or increasing) to reactive (i.e., decreasing dissolved concentrations)  
87 behaviour of dissolved aluminum in several pit lakes has been observed to occur at pH~4.0  
88 (Sánchez-España et al., 2011), which is consistent with previous observations in tailings pore  
89 waters (Blowes et al., 2003) as well as in acid sulfate soils (Mosley et al., 2014a,b).

90 The behaviour of Al at pH<4.0 has been controversial (see discussions in Bigham and  
91 Nordstrom, 2000, Jones et al., 2011, and Sánchez-España et al., 2011). Previous studies  
92 proposed that jurbanite ( $\text{AlSO}_4\text{OH}\cdot 5\text{H}_2\text{O}$ ) could control the solubility of Al in acid sulfate  
93 soils and solutions (e.g., Van Bremen, 1973; Karathanasis et al., 1988; Zhu and Burden,  
94 2001). This mineral has also been proposed to control the concentration of Al in pit lakes  
95 (Eary, 1999; Eary and Castendyk, 2013). However, in most occasions the possible role of  
96 jurbanite has been suggested solely by geochemical modelling, while the presence of this  
97 mineral in the studied systems has never been demonstrated. Jurbanite has only been observed  
98 as a typical efflorescent mineral (Anthony and McLean, 1976) and this mineral has probably  
99 no role on the control of Al solubility in acidic waters.

100 The aim of this study was to expand the current knowledge on the mobility of Al at low  
101 pH. For our research, we selected several pit lakes of the IPB in SW Spain. These artificial  
102 water bodies constitute field-scale laboratories of incomparable value, hosting highly acidic  
103 and Al-rich water over long periods of time (often decades). Furthermore, the wide  
104 compositional spectrum found in these lakes (pH=2.2-4.5, [Al]=2-20 mM,  $[\text{SO}_4^{2-}]$ =0.02-0.3

105 M) provides a suitable window for the study of aluminum solubility under typical conditions  
106 of AMD and many acidic soils and sediment pore waters. In this first paper of a two-part  
107 series, we focus on the behaviour of Al at the lowest pH range (pH<4.0) where Al  
108 precipitation is both kinetically and thermodynamically unfavourable and where this metal is  
109 often considered to behave conservatively.

110 With this purpose, we studied mineral colloids found at different depths (ranging from  
111 near-surface to 100 m) in the water column of the highly acidic (pH 2.2-3.1) San Telmo pit  
112 lake. To expand the pH window of our study, we also include observations and analyses  
113 obtained on schwertmannite particles and jarosite crystals formed at slightly higher pH (up to  
114 4.3) in the nearby pit lakes of Cueva de la Mora and Herrerías. Further limnological,  
115 hydrogeochemical and microbiological information about these lakes can be found in recent  
116 studies (*e.g.*, Sánchez-España et al., 2009, 2011, 2012, 2013, 2014; Díez-Ercilla et al., 2009,  
117 2014; Wendt-Potthoff et al., 2012; Falagán et al., 2014).

## 118 2. MATERIALS AND METHODS

### 119 2.1. Environmental framework

120 The San Telmo pit lake (Fig. 1) was formed by flooding of a former, large open pit which  
121 was mined until 1990 for the recovery of Cu, Zn and Pb from massive sulphide ore  
122 mineralization. The ore deposit was hosted in hydrothermally altered volcanic rocks (mainly  
123 rhyolites) associated to clastic sedimentary rocks (shales) of Carboniferous age. Today this  
124 130 m-deep lake is meromictic and comprises an upper mixolimnion of around 29 m depth  
125 (which seasonally mixes and experiences winter overturn) and a deep monimolimnion which  
126 is perennially isolated from the atmospheric conditions. However, unlike most other  
127 meromictic pit lakes of the IPB, where Fe is solely present as Fe(II) in the deep waters, the  
128 redox potential stays relatively high (ORP~420-460 mV) even in the deep, anoxic  
129 monimolimnion (Fig. 1d), and Fe(III) is present throughout the whole water column. This is  
130 an unusual feature which makes this lake an outstanding “geochemical reactor” with more  
131 than 8 Mm<sup>3</sup> of Fe(III)-precipitating acidic water. The pH is strongly buffered by the  
132 precipitation of Fe(III) in the form of either schwertmannite and/or jarosite, though  
133 schwertmannite is metastable and tends to transform to jarosite and/or goethite during aging  
134 (Bigham et al., 1996; Sanchez-España et al., 2011, 2012). Seasonal fluctuations of pH  
135 throughout the year (Fig. 1e) result from the combined effects of diverse geochemical and

136 hydrological factors, including: (i) volume and composition of inflowing waters –mostly  
137 AMD seeping from adjacent waste piles, but also groundwater and direct meteoric  
138 precipitation–, (ii) evapo-concentration of solutes during the dry summer season, (iii)  
139 microbially catalysed oxidation of Fe(II), (iv) microbially mediated reduction of Fe(III) and  
140  $\text{SO}_4^{2-}$ , and (v) water/rock interaction, including the dissolution of acidity-releasing (*e.g.*,  
141 pyrite) and acidity-consuming (*e.g.*, carbonate, silicate) minerals.

142 There is a marked increase in conductivity and sulfate concentration below the redoxcline  
143 (Sanchez-España et al., 2012). Most metals behave conservatively and show a vertical pattern  
144 of nearly constant (*e.g.*,  $\text{SiO}_2$ , Cl) or increasing (*e.g.*, Ca, Mg, Mn, Cu, Zn, Cd, Co, Ni)  
145 concentration with depth (Table 1; *see also Fig. EA1 in electronic annex*). In this context, both  
146 Fe(III) and Al display an opposite trend, showing clear signs of removal by precipitation  
147 and/or sorption reactions (Table 1; Fig. 2a). Furthermore, some other trace elements like K  
148 (Fig. 2b), and toxic metal(oid)s like As, Cr, Mo and Pb (Table 1; Fig. 2c-d), show a similar  
149 decreasing trend which is likely linked to the mobility of both iron and aluminium, as  
150 discussed below.

## 151 2.2. Field work and sampling

152 The waters and solids studied in this work were obtained during different field campaigns  
153 conducted between 2006 and 2015. We used densely coated membrane filters (0.1 and 0.45  
154  $\mu\text{m}$  pore size; Millipore®) to study the chemical and mineralogical nature of colloidal  
155 precipitates existing at different depths in the lakes. These filters were obtained by filtering  
156 250 mL of water collected from discrete depths. The filters were washed by subsequent  
157 filtration of 50-100 mL of ultra-pure deionized water (MilliQ), stored at 4 °C and used for  
158 mineralogical identification and chemical characterization. In addition, we also used sediment  
159 traps installed in the San Telmo pit lake at depths of 30, 40 and 100 m. These traps were  
160 sampled every 3-4 months during this time interval. All traps were attached to 6 mm-thick  
161 mooring lines anchored to the lake bottom. These PVC traps (supplied by Hydro-Bios; GmbH,  
162 Kiel-Holtenau, Germany) had a collection area of 150  $\text{cm}^2$ , and a cylinder length of 56 cm  
163 within a 250 mL polyethylene bottle. The sediment traps collected a significant amount of  
164 chemical precipitates formed *in situ* at depth, in addition to fine-grained sediments that settled  
165 from above. Immediately after collection, these sediments were carefully washed by  
166 submersion in consecutive MilliQ water baths (without sonication or shaking), dried at room  
167 temperature and stored at ambient conditions until microscopic analysis.

168 Field measurements (pH, ORP, T, O<sub>2</sub>) were taken with a Hydrolab MS5 multi-parametric  
169 probe (Hach®, Loveland, CO, USA), after calibration with fresh standards. Water samples for  
170 chemical analyses were taken from different depths with a Van Dorn® sampling bottle (KC  
171 Denmark). All samples were filtered on site with 0.45 µm nitrocellulose membrane filters  
172 (Millipore®), stored in 125 mL-polyethylene bottles, acidified with HNO<sub>3</sub> (1 M), and  
173 preserved on ice in coolers during transport.

### 174 **2.3. Laboratory analyses of waters**

175 Water samples were analyzed by atomic absorption spectrometry (AAS; major cations),  
176 inductively coupled plasma-atomic emission spectrometry (ICP-AES; S as SO<sub>4</sub><sup>2-</sup>), and  
177 inductively coupled plasma-mass spectrometry (ICP-MS; trace metals) using Varian  
178 SpectrAA 220 FS, Varian Vista MPX and Agilent 7500ce instruments, respectively. For  
179 simplification, only some of these elements will be discussed in this work. The detection limit  
180 for major cations was <1 mg/L. The detection limits for trace elements was usually between  
181 0.4 µg/L (e.g., As) to 10 µg/L (e.g., Pb).

### 182 **2.4. Chemical and mineralogical identification of bulk Fe(III)-Al precipitates**

183 The aluminous and ferric precipitates were initially characterized by inductively coupled  
184 plasma atomic emission spectroscopy (ICP-AES) after previous acid digestion with HF,  
185 HClO<sub>4</sub>, HNO<sub>3</sub>, and HCl for the determination of major cations (Al, Fe) and trace metal(oid)s  
186 (including As, Pb, Cu and Zn). These solids were also mineralogically identified by powder  
187 XRD on a PANalytical X'Pert Pro diffractometer (X'Pert software with ICDD database), with  
188 Cu K $\alpha$  radiation (40 kV, 40 mA), graphite monochromator and automatic slit. Cryogenic <sup>57</sup>Fe  
189 transmission Mössbauer spectroscopy was also carried out on selected ferric precipitates from  
190 the San Telmo pit lake at a temperature of 5 K. Solid material was scraped off filter paper and  
191 sealed between two pieces of 5 Mil Kapton tape and mounted on the sample holder. The <sup>57</sup>Co  
192 radioactive source (~50 mCi) resided in an R<sub>h</sub>-matrix at room temperature during data  
193 collection. Spectral fitting was done using Recoil software (University of Ottawa, Canada)  
194 using the Voigt-based model (Rancourt and Ping, 1991). Data were calibrated against an  $\alpha$ -  
195 <sup>57</sup>Fe foil collected at room temperature.

### 196 **2.5. Scanning and transmission electron microscopy characterization of Fe(III)-Al** 197 **particles/crystals**

198 The aluminous and ferric precipitates naturally formed in the lakes were initially studied  
199 by scanning electron microscopy (SEM) coupled with an energy-dispersive X-ray  
200 spectrometer (EDS), and transmission electron microscopy (TEM) equipped with EDS. These  
201 studies were conducted at the SGIker Advanced Research Facilities (UPV/EHU).  
202 Compositional analyses were carried out on carbon-coated samples by EDS on a JSM-7000F  
203 field emission scanning electron microscope (JEOL) working at 20 kV. These analyses  
204 provided valuable information of mineral chemistry on a sub-micron analyses volume. Among  
205 hundreds of analyses conducted in the precipitates from the lakes, we only considered those  
206 showing no evidence of fine-grained detrital silicate contamination. TEM images and selected  
207 area electron diffraction (SAED) analyses of Fe(III) and Al compounds were performed on a  
208 Philips CM200 transmission electron microscope after placing 2-3 drops of an ethanol  
209 suspension with the fine precipitates on a porous carbon film supported by a Cu grid.

210 Ultra high-resolution imaging and compositional analysis was performed using a double  
211 aberration-corrected FEI Titan<sup>3</sup> G2 60-300 scanning/transmission electron microscope  
212 (S/TEM) at the Materials Characterization Laboratory (MCL) of the Materials Research  
213 Institute (MRI) in The Pennsylvania State University. This microscope has sub-Ångstrom  
214 resolution and can be used to determine both crystal structure and elemental composition over  
215 large areas of the sample. EDS maps were acquired using FEI's ChemiSTEM technology,  
216 which combines a high brightness Schottky field emission gun with four Super-X silicon drift  
217 x-ray detectors. This EDS system produces very high x-ray count rates, which allowed  
218 crystals to be compositionally mapped at an average time of 5 minutes per map and a beam  
219 current of approximately 0.7 nA. Samples were sonicated in an ethanol bath and then placed  
220 on lacey carbon-coated Cu grids before being inserted into the microscope. The samples were  
221 studied using both 80 and 200 kV for HAADF (high angle annular dark field) STEM, which  
222 provides high resolution images containing primarily mass contrast, and EDS elemental  
223 mapping of crystals and particles with diameters comprised between 200 nm and 1  $\mu\text{m}$ . EDS  
224 mapping at low magnification (e.g., 5,000 $\times$ ), either displayed independently or combined with  
225 the HAADF-STEM image, is useful for rapid identification of minerals over a relatively large  
226 selected area. Further analyses of selected crystals at higher magnification (up to 80,000 $\times$ )  
227 permitted a more thorough investigation of Al and Fe(III) distribution via detailed crystal EDS  
228 mapping and transversal (through-crystal) EDS line scans. The spectra obtained by EDS  
229 analysis was quantified by the Cliff-Lorimer ratio method to obtain the relative concentrations



230 between the elements from the EDS peak intensities after performing a background subtraction  
231 and peak deconvolution using the Bruker Espirit software.

232 We also obtained structural information on selected alunite and jarosite crystals by high  
233 resolution transmission electron microscopy (HRTEM), including d-spacings measured on the  
234 HRTEM images and/or between diffraction spots and rings in the SAED patterns.

235 EDS has long been used in geology to aid in the microscopic study of minerals and  
236 compounds (Friel et al., 2003; Garratt-Reed and Bell, 2003, and references therein) and  
237 recently has greatly improved the performance and reliability for chemical characterization of  
238 materials after the incorporation of silicon drift detectors to the modern scanning and  
239 transmission electron microscopes (Newbury and Ritchie, 2013a,b). Our analyses were  
240 calibrated with different minerals of known mineralogy which were used as internal standards.  
241 We then compared the results (after adequate peak verification and background correction) to  
242 the corresponding ideal compositions. A selection of “pure specimen” analyses (conducted on  
243 minerals almost free of trace elements) is provided in the supplemental material (*Table EA1*).  
244 In general, a close similarity was found for the EDS analyses of pure schwertmannite, K-  
245 jarosite, H<sub>3</sub>O-jarosite, hydrobasaluminite, alunite and gypsum crystals with respect to their  
246 ideal compositions, as well as with analyses previously obtained by other techniques. The  
247 variability in the concentration of elements like Al, Fe, K, O and S in jarosites and alunites  
248 were thus considered natural chemical trends resulting from K<sup>+</sup>/H<sub>3</sub>O<sup>+</sup> or Fe/Al substitution.  
249 The K<sup>+</sup>/H<sub>3</sub>O<sup>+</sup> molar ratio of the analysed jarosites and alunites was calculated from the EDS  
250 analyses based on the theoretical stoichiometry (one mol of alkalies per formula unit) and  
251 inferring the hydronium content as [H<sub>3</sub>O<sup>+</sup>]=(1-K<sup>+</sup>-Na<sup>+</sup>). Taking into account the small size of  
252 these precipitates (usually below 1 μm), STEM-EDS analyses provide the most reliable  
253 compositions and better define the chemistry of the whole crystal.

## 254 **2.6. Water/mineral equilibrium calculations**

255 We used PHREEQC (Version 3.0.5-7748; Parkhurst and Appelo, 2013) to calculate the  
256 saturation index (SI) of Fe(III) and Al minerals, including alunite, gibbsite, goethite, H<sub>3</sub>O-  
257 jarosite, K-jarosite, and schwertmannite, in the pit lake waters. All calculations were made  
258 using the WATEQ4F thermodynamic database (Ball and Nordstrom, 1991), previously  
259 modified with solubility product constants for K<sup>+</sup>- and H<sub>3</sub>O<sup>+</sup>-jarosite end-members (taken  
260 from Alpers et al., 1989 and Baron and Palmer, 1996) and schwertmannite (taken from

261 Bigham et al., 1996). For these calculations, we considered physico-chemical parameters  
262 measured on site (pH, ORP, T, O<sub>2</sub>) and included all the cationic and anionic species in  
263 solution. Ionic activities were calculated with the Davies equation (Davies, 1962). The ionic  
264 strength of the analyzed waters was around 0.1-0.2 molar. The pressure gradient across the  
265 water column (1-10 bar) is small in absolute terms, and its effect on water/mineral equilibrium  
266 was not considered in the calculations. We did consider, however, the variation of water  
267 density between different depths, as this parameter affects the molal proportions calculated  
268 from solute concentrations introduced as mg L<sup>-1</sup> or µg L<sup>-1</sup>. Water density was calculated based  
269 on temperature and solute concentration using the RHOMV 2.0 numerical program (Boehrer  
270 et al., 2010; available at [www.ufz.de](http://www.ufz.de)).

### 271 3. RESULTS

#### 272 3.1. General chemical and microscopic features

273 Based on XRD, mineral precipitates found in the deep (i.e., >35 m) anoxic waters of the  
274 San Telmo acid pit lake (pH<3.2) were dominated by jarosite along with trace amounts of  
275 schwertmannite and quartz. The predominance of Fe(III) [and absence of Fe(II)] in these  
276 minerals was confirmed by Mössbauer spectroscopy (*Fig. EA3 in electronic annex*). Even  
277 though the sample contained trace schwertmannite, the Mössbauer spectra were well fit to a  
278 reference jarosite spectrum, consistent with XRD results (not shown). Whole-sample chemical  
279 analyses carried out by ICP-AES on these solids yielded total iron concentrations in the range  
280 of 25.9-27.4 wt.% Fe which are lower than theoretical values of ideal K- or H<sub>3</sub>O-jarosite  
281 (33.9-35.5 wt.%) (*Table EA2 in electronic annex*). These precipitates also contained  
282 significant Al (3.9-4.3 wt.%) in addition to trace components such as As, Pb, Cu and Zn at  
283 concentrations commonly exceeding 1,000 mg/kg (*Table EA2*). Detailed SEM and STEM  
284 studies have revealed that these trace constituents were incorporated into both jarosite and  
285 schwertmannite through adsorption and/or ionic substitution, as discussed below.

286 The microscopic examination of these particles by SEM and TEM also showed a mineral  
287 assemblage dominated by jarosite and coexisting schwertmannite at all depths, though  
288 subordinate alunite was also observed (*Fig. 3*). Schwertmannite displays a typical *hedge-hog*  
289 morphology with nanometric whiskers that grew radially and formed pseudo-spherical  
290 aggregates (*Fig. 3a*), while jarosite and alunite were usually present as idiomorphic crystals  
291 with pseudo-cubic to rhombic habit and diameters ranging from 200 nm to 1 µm. Some

292 jarosite crystals displayed a clear concentric zoning which indicates chemical heterogeneity  
293 during successive stages of crystal growth (Fig. 3c-d). Textural and compositional features of  
294 these coexisting minerals suggest that many jarosite crystals actually resulted by  
295 transformation of previously formed schwertmannite, though direct precipitation of jarosite  
296 from the parent fluid also takes place at  $\text{pH} < 2.5$  (Sanchez-España et al., 2012).

297 Jarosite crystals often showed an important potassium deficit suggestive of hydronium  
298 substitution (Fig. 3a-b), which has led to their classification as hydronian ( $\text{H}_3\text{O}^-$ ) jarosites  
299 (Sánchez-España et al., 2012). Further, SEM spot-size EDS analyses conducted on selected  
300 areas of schwertmannite particles and jarosite crystals indicated that these two minerals may  
301 contain significant amounts of adsorbed and/or coprecipitated elements such as Al, Mg and Si  
302 (Fig. 3a-b). Most notably, the variable and locally high concentration of Al found in  
303 schwertmannite and jarosite was also indicative of an important degree of Fe(III)-Al co-  
304 precipitation by either Al adsorption (likely dominant in schwertmannite) and/or Al/Fe  
305 substitution (prevailing in jarosite). To confirm this possibility, we used STEM to conduct  
306 spatially-resolved, whole-crystal elemental mapping in schwertmannite particles and jarosite  
307 crystals.

### 308 3.2. Schwertmannite chemistry

309 The STEM-EDS images obtained for schwertmannite particles formed in the studied pit  
310 lakes (San Telmo, Cueva de la Mora and Herrerías) revealed a variable distribution of Al.  
311 Some aggregates showed a homogenous distribution of Al throughout the particles, nearly  
312 identical to that of major components such as Fe and S (Fig. 4).

313 In any case, the concentration of Al in schwertmannite was variable and seemed to be pH-  
314 dependent (Table 2; Fig. 5). Among the chemically homogenous schwertmannite particles,  
315 those found at higher pH ( $\sim 4.0$ ) had higher Al content (2.56-3.37 at.%,  $\text{Fe}/[\text{Fe}+\text{Al}] = 0.89-$   
316  $0.90$ ), while those formed at lower pH (2.6-3.5) showed lower Al content ( $\text{Al} < 1.2$  at.%,  
317  $\text{Fe}/[\text{Fe}+\text{Al}] = 0.95-1.00$ ).

318 At higher pH ( $> 4.0$ ), Al tended to be more heterogeneously distributed and chemical  
319 partitioning was clearly visible, suggesting a distinct mechanism of entrapment. Nano-scale  
320 elemental mapping and chemical through-particle scanning of schwertmannite particles found  
321 in the deep part of the Herrerías pit lake (pH 4.3) showed Al to be especially concentrated in  
322 outer areas adjacent to the needles (Fig. 6). The example illustrated in Fig. 6 (*see also Fig.*

323 *EA4 in electronic annex*) shows the strong chemical contrast between the core and needles in  
324 schwertmannite particles with typical *pincushion* morphology. Even though Al was always  
325 present in the inner zones at significant concentrations (*e.g.*, 8.2-8.4 at.%), Al was far more  
326 abundant in outer zones (*e.g.*, 15.8-17.1 at.% Al) (Table 2; Fig. 6b), where it seemed to  
327 surround the whiskers (Fig. 6c-d).

328 While the inner zones (with Fe/S=6.2-6.5 and Fe/[Fe+Al]=0.77-0.79) may still represent a  
329 non-ideal schwertmannite with a higher degree of Al/Fe substitution due to a higher pH, the  
330 outer zones have a very distinct composition (with Fe/S=2.6-2.8 and Fe/[Fe+Al]=0.41-0.48)  
331 and more likely represent an Al shell formed by adsorption on pre-existing schwertmannite  
332 (Table 2; Fig. 5a).

333 Geochemical modelling suggests that, within the pH range of most acidic mine waters  
334 (2.0-4.5), dissolved Al is mostly present as  $\text{AlSO}_4^+$  and  $\text{Al}(\text{SO}_4)_2^-$  complexes, along with  
335 minor  $\text{Al}^{3+}$  (Fig. 5b; Sanchez-España, 2007). An important inflection point exists around pH  
336 3.3 and marks the dominance of the cationic  $\text{AlSO}_4^+$  (dominant at  $\text{pH}<3.3$ ) versus the anionic  
337  $\text{Al}(\text{SO}_4)_2^-$  (dominant at  $\text{pH}>3.3$ ).

### 338 3.3. Jarosite and alunite chemistry

339 The presence of relatively pure Al minerals (*e.g.*, alunite, gibbsite) could not be detected  
340 by XRD. However, the use of chemical mapping with STEM allowed us to identify both  
341 jarosite and alunite nanocrystals, and to measure the distribution of Fe and Al in these crystals  
342 (Figs. 7-9; *see also Figs. EA5 and EA6 in the electronic annex*).

343 Chemical mapping of Fe(III) and Al showed that nearly pure (end-member) jarosite could  
344 coexist with both end-member alunite and concentrically zoned Al-Fe crystals. These crystals  
345 were found to be composed of a jarositic rim (250-400 nm thickness) surrounding an Al-rich  
346 core with composition intermediate between alunite and jarosite (Figs. 7-8). In some of these  
347 zoned crystals, the Al content decreased progressively from the crystal center to the outer  
348 rims, suggesting a replacement process defined by the isomorphic transformation/substitution  
349 of previously formed alunite to more stable jarosite during ageing (Fig. 7). In other cases,  
350 however, the trends of Fe(III) and Al distribution within the crystals were not progressive and  
351 more likely reflected distinct stages of crystal growth (Fig. 8). Such distinct stages could have  
352 occurred during slightly varying geochemical conditions (*e.g.*, pH) in the parent solution, in

353 accordance with previous findings in other jarosite growth zonings described in the literature  
354 (e.g., Papike et al., 2006; Burger et al., 2009).

355 Chemically homogeneous crystals of nearly pure alunite with very low content of Fe(III)  
356 were also observed in these samples (Fig. 9), though they were far less abundant than jarosite  
357 at all depths, consistent with the inability to detect alunite by XRD.

358 Electron diffraction and high resolution images of selected areas of these crystals helped  
359 confirm the mineralogical nature of these two phases (Fig. 10). Although we could not obtain  
360 the full set of diagnostic d-spacings, many of the Fe-rich crystals yielded d-spacings matching  
361 with those of K-jarosite and/or H<sub>3</sub>O-jarosite (e.g., Fig. 10a). In the case of alunite, the  
362 diffractometric identification was less clear since the crystals apparently displayed a slightly  
363 lower crystallinity with lesser rings and spots. However, we could find a few d-spacings  
364 corresponding to alunite in the Al-rich crystals (e.g., Fig. 10b).

365 Chemical quantification of both chemically homogeneous and concentrically zoned  
366 crystals of jarosite and alunite indicated highly variable concentrations of Al and Fe  
367 ( $\text{Fe}/[\text{Fe}+\text{Al}]_{\text{m}}=0.54\text{-}0.99$  in jarosites and  $\text{Fe}/[\text{Fe}+\text{Al}]_{\text{m}}=0.07\text{-}0.37$  in alunites; Table 3, Fig. 11;  
368 see also Table EA3 in electronic annex). The potassium content was also highly variable  
369 ( $\text{H}_3\text{O}^+/[\text{H}_3\text{O}^++\text{K}^+]_{\text{m}}=0.01\text{-}0.86$  in jarosites, and  $\text{H}_3\text{O}^+/[\text{H}_3\text{O}^++\text{K}^+]_{\text{m}}=0.01\text{-}0.55$  in alunites) and  
370 indicated variable H<sub>3</sub>O<sup>+</sup>/K<sup>+</sup> substitution in both minerals.

371 Saturation index (SI) calculations for a number of Fe(III) and pure Al phases in the San  
372 Telmo pit lake help explain the equilibrium conditions of jarosite and alunite stability (Figs.  
373 12-13). The evolution of SI with depth is shown in Fig. 12, while the evolution of these  
374 indices with pH is provided in Fig. 13a. The plots of Fig. 12 reflect two different scenarios  
375 which illustrate the situation at the beginning of summer (June 2006, pH=2.4-2.6; Fig. 12a)  
376 and at the beginning of spring (April 2011; pH=2.8-3.0; Fig. 12b). In April 2011, the  
377 mixolimnetic waters were strongly oversaturated with respect to schwertmannite, and this  
378 mineral is commonly the first and most abundant product of Fe<sup>III</sup> precipitation in the lake.  
379 Below the redoxcline, however, schwertmannite is no longer stable (SI<0) and tends to  
380 transform to jarosite and/or goethite (Sanchez-España et al., 2012). The low pH of the lake  
381 favours the substitution of H<sub>3</sub>O<sup>+</sup> for K<sup>+</sup> in the alkali site and the formation of hydronian  
382 jarosite, which is common in very acidic systems (e.g., Alpers et al., 1989; Dutrizac and  
383 Jambor, 2000; Stoffregen et al., 2000; Jamieson et al., 2005). On the other hand, Al minerals

384 (exemplified here by alunite and gibbsite) are notably undersaturated at all depths and in the  
385 two periods studied (Fig. 12), consistent with the relative scarcity of alunite and the absence of  
386 gibbsite. However, the finding of alunite crystals indicates that this mineral was likely formed  
387 under slightly different geochemical conditions (*e.g.*, higher pH) with respect to those  
388 measured in the parent fluid at the moment of sampling.

389 The calculated saturation indices plotted against pH help define pH limits and stability  
390 fields for the studied minerals (Fig. 13a). This plot indicates precipitation of schwertmannite  
391 at  $\text{pH} > 2.5$ , and precipitation of  $\text{H}_3\text{O}$ -jarosite at  $\text{pH} > 2.4$  for the geochemical conditions found  
392 in the San Telmo pit lake. K-jarosite is oversaturated in every case and its precipitation is  
393 therefore feasible in the whole pH window (2.0-3.5). Alunite is always undersaturated, as  
394 already stated, and extrapolation of the available data indicate that it could precipitate at a pH  
395 around 3.3 (Fig. 13a). This pH is only slightly higher than the highest pH value measured in  
396 the San Telmo pit lake (3.1) and may well have existed at certain periods or in restricted  
397 micro-environments such as in the surroundings of microbial cells.

398 The pH limits obtained in Fig. 13a allowed us to establish a geochemical model for the  
399 Fe(III)-Al system with stability fields for the most relevant Fe(III) and Al minerals formed in  
400 AMD systems (Fig. 13b). We have considered the electronic potential (*pe*) to discriminate  
401 between Fe(II)- and Fe(III)-dominated systems, as the presence of Fe(III) may strongly  
402 influence the behaviour of dissolved Al at low pH. We selected a limit of  $\text{pe} \sim 6.7$  ( $\text{Eh} \sim 400$   
403 mV) to separate both fields based on previous studies (*e.g.*, Sanchez-España et al., 2005; Diez-  
404 Ercilla et al., 2014). We have also enlarged the pH window to pH 7 to include phases stable  
405 under more neutral conditions. The diagram was produced considering the Fe-Al-K-S-O-H  
406 system for relatively dilute conditions ( $a_{\text{SO}_4} = 10^{-2}$ - $10^{-1}$ ,  $a_{\text{Fe}} = a_{\text{Al}} = 10^{-3}$ ,  $a_{\text{K}} = 10^{-5}$ ). Therefore,  
407 phases forming in more concentrated solutions (*e.g.*, evaporative brines), such as alunogen, or  
408 phases incorporating other cations (*e.g.*, natro-jarosite) were not considered.

## 409 4. DISCUSSION

### 410 4.1. Incorporation of Al into schwertmannite

411 It is well known that  $\text{Al}^{3+}$  can substitute isomorphously for  $\text{Fe}^{3+}$  in iron oxides, and may  
412 occupy up to a third of the structural sites of  $\text{Fe}^{3+}$  in minerals such as goethite (Cornell and  
413 Schwertmann, 2003). Although schwertmannite is a well-known adsorbent for many toxic  
414 trace elements, including As and Cr (*e.g.*, Regenspurg and Peiffer, 2005; Sanchez-España et

415 al., 2005, 2006; Burton et al., 2009), the extent and mechanisms of Al incorporation into this  
416 mineral have not been studied in detail.

417 The homogeneous distribution of Al throughout many schwertmannite particles (Fig. 4d)  
418 points to a likely incorporation of Al into schwertmannite, either by adsorption and/or through  
419 inclusion and later entrapment during precipitation and subsequent mineral growth. The  
420 adsorption of Al on schwertmannite is supported by the geochemical environment in which  
421 the schwertmannite particles were found. The monimolimnion in the Herrerías pit lake is  
422 anoxic and lacks any Fe(III) in solution, so that the schwertmannite particles were probably  
423 formed in the upper, oxidizing layer and were subsequently transported downwards by  
424 settling. The interaction of Al ionic complexes in solution with the schwertmannite particles  
425 during settling probably led to this strong Al adsorption on the high surface area whiskers.  
426 This adsorption may have been favoured by the electrostatic attraction between the positively  
427 charged surfaces of schwertmannite particles (with  $\text{pH}_{\text{ZPC}}=6.6-7.1$ ; Regenspurg, 2002) and  
428 negatively charged species like  $\text{Al}(\text{SO}_4)_2^-$ . The feasibility for Al sorption onto pre-existing  
429 Fe(III) phases in the pH range covered in this study is supported by geochemical modelling  
430 (Fig. 5b). In contrast, very limited or negligible Al sorption should be expected at  $\text{pH}<2.5$   
431 (either for absence of suitable Fe(III) adsorbent phase or due to dominance of  $\text{Al}^{3+}$  and  $\text{AlSO}_4^+$   
432 species) and at  $\text{pH}>4.5$  (where Al removal will occur chiefly by precipitation as Al-  
433 oxyhydroxysulfate and/or Al-hydroxide minerals).

434 The kinetics of Al sorption and incorporation into schwertmannite is unknown, but the  
435 long residence time of the deep waters in these lakes and the slow settling velocity of the fine-  
436 grained schwertmannite particles are both hydraulic factors that should enhance adsorption.

437 The possibility of structural substitution of  $\text{Al}^{3+}$  for  $\text{Fe}^{3+}$  cannot be ruled out, though the  
438 existence of Al-substituted schwertmannites cannot be demonstrated in the present work. This  
439 question will require further in-depth nano-structural and crystallographic studies.

#### 440 **4.2. The jarosite-alunite solid solution**

441 The inverse correlation between Fe and Al concentration in the analysed jarosite and  
442 alunite crystals likely reflects Fe/Al substitution and a solid solution series between the two  
443 end-members (Fig. 11a). Considering only the data set from San Telmo ( $R^2=0.97$ ), the Fe-Al  
444 analytical trend parallels the ideal solid solution trend. Additional data from other pit lakes  
445 (Cueva de la Mora, Herrerías) introduce some scatter, but do not reveal a bimodal distribution.

446 However, when considering atomic proportions plotted in a binary diagram of  $\text{Fe}/[\text{Fe}+\text{Al}]_m$  vs.  
447  $\text{H}_3\text{O}^+/[\text{H}_3\text{O}^++\text{K}^+]_m$ , it seems that two populations do exist (Fig. 11b). The analytical points can  
448 be grouped into either Fe-rich alunites or Al-rich, hydronian jarosites, with few points lying in  
449 between. This plot reveals a double solid solution series where  $\text{Fe}^{3+}/\text{Al}^{3+}$  substitution and  
450  $\text{K}^+/\text{H}_3\text{O}^+$  substitution are coupled and where the hydronian substitution in the jarosites is  
451 notably more important than in alunites. It is difficult to ascertain whether the scarcity of  
452 intermediate points reflects a crystallographically controlled miscibility gap or simply reflects  
453 the very singular geochemical conditions enabling the formation of transitional compositions  
454 with nearly equivalent molar proportions of Fe and Al. The existence of a few intermediate  
455 compositions may suggest that these compositions are structurally feasible but waters with  
456 comparable concentrations of Al and Fe(III) are an underrepresented geochemical condition  
457 required for their formation.

458 The complete solid solution between alunite and jarosite has been experimentally  
459 synthesized in the laboratory (Brophy et al., 1962; Härtig et al., 1984), however the  
460 occurrence of this solid solution has rarely been found in nature (Brophy et al., 1962; Alpers et  
461 al., 1989, 1992). Compositions intermediate between jarosite and alunite end members (i.e.,  
462 those with comparable molar proportions of Fe and Al in the octahedral (B) site) are very  
463 uncommon in the literature, which has led some authors hypothesize about the possible  
464 existence of a miscibility gap in this solid solution. It has been argued that such a miscibility  
465 gap is unlikely (*e.g.*, see discussions in Alpers et al., 1989, Stoffregen et al., 2000, Dutrizac  
466 and Jambor, 2000, or Papike et al., 2006), so that the apparent scarcity of intermediate  
467 compositions would be more likely a result of the physical separation of  $\text{Fe}^{3+}$  from  $\text{Al}^{3+}$  in  
468 natural waters due to their very different first hydrolysis constants ( $\text{pK}_1 \text{Fe}^{3+}=2.19$ ,  $\text{pK}_1$   
469  $\text{Al}^{3+}=4.99$ ; Nordstrom and Ball, 1986). Alunite formation seems to be enhanced in  
470 environments where ferric iron is absent or not abundant (Keith et al., 1979; Alpers et al.,  
471 1989). The results obtained in this study suggest that both alunite and jarosite may form and  
472 coexist in the same aqueous environment, provided that suitable geochemical conditions are  
473 met. These conditions ideally include comparable molar concentrations of Al and Fe(III), and  
474 a pH window enabling the formation of both minerals (see discussion below). The formation  
475 of Fe(III)-containing alunite and Al-rich jarosite in the monimolimnion of the San Telmo pit  
476 lake appears to be enhanced by several factors, most notably the higher molar concentration of  
477 aluminum with respect to that of ferric iron (5-7 mM  $\text{Al}^{3+}$  vs. 2-3 mM  $\text{Fe}^{\text{III}}$ ; Fig. 2a). If it



478 exists, the compositional gap in this solid solution is probably significantly smaller than what  
479 has been traditionally considered.

480 Furthermore, the presence of isolated Fe-containing alunite crystals not associated with  
481 schwertmannite suggests that these solids have directly precipitated from the parent solution  
482 by incorporation of  $\text{Al}^{3+}$  for  $\text{Fe}^{3+}$  into the mineral structure during crystallization. The zoned  
483 crystals with distinct alunite and jarosite domains likely reflect the metastability of the former  
484 with respect to the later, so that many analyses in the plots of Fig. 11 may well reflect different  
485 stages of an alunite to jarosite isomorphic replacement with different degrees of completion.

486 Although the hydronium substitution was always much more important in jarosite, it is  
487 worth to note that we have also identified an important presence of  $\text{H}_3\text{O}^+$  in the alkali site in  
488 alunites. This  $\text{H}_3\text{O}^+/\text{K}^+$  substitution has been rarely demonstrated in nature, but has been  
489 inferred in both synthetic and low-temperature natural alunites because of a deficiency in  
490 alkalis and an excess of water compared with the stoichiometric composition (e.g., Ripmeester  
491 et al., 1986; Stoffregen et al., 2000).

492 The analyzed jarosites and alunites could also incorporate small quantities of Si and Mg  
493 (0.4-6% Si; 0.5-2.6% Mg; see Table EA3 in electronic annex), though these two elements are  
494 considered to be adsorbed rather than structural. Silicon is included in the formula of some  
495 rare minerals of the alunite supergroup (e.g., waylandite, eylettersite), but the formation of  
496 these minerals requires the presence of  $\text{SiO}_4^{4-}$  which is only possible under highly basic  
497 conditions. Under acidic conditions, analytically determined silicon is most likely present as  
498 adsorbed complex anions or as amorphous silica gels and not as solid-solution Si (Dutrizac  
499 and Jambor, 2000).

### 500 4.3. Water/mineral equilibrium

501 According to the diagram of Fig. 13b, alunite would be only stable over a narrow window  
502 of pH (3.3-4.0). Above pH 4.0, the precipitation of hydrobasaluminite and/or felsobányaite  
503 would be kinetically favoured (Sanchez-España et al., 2011) and these minerals would be the  
504 main Al sink up to a pH around 6.0. Above pH 6.0, Al-OH complexes become more abundant  
505 than Al-SO<sub>4</sub> complexes (Fig. 5b; Sanchez-España, 2007) and gibbsite formation is favoured.  
506 The formation of gibbsite and hydrobasaluminite has been observed in the studied lakes with  
507 pH>4.0, and will be presented in another study.

508 Below pH 3.3, the behaviour of Al is strongly dependent on redox potential. At  $p_e > 6.7$ ,  
509 the presence of Fe(III) enhances the formation of different phases, depending on pH  
510 conditions. K-jarosite is stable above pH 2.0, while schwertmannite is both kinetically and  
511 thermodynamically favoured at  $pH > 2.5-2.6$  (Fig. 13a-b). The stability field of hydronian  
512 jarosite is limited by its apparent instability below pH 2.3 and the aforementioned faster  
513 kinetics of schwertmannite precipitation above pH 2.5. Under reducing conditions with low  
514 redox potential, however, the absence of Fe(III) solids to adsorb and/or entrap Al would lead  
515 this metal to stay in solution in different ionic forms ( $AlSO_4^+$ ,  $Al(SO_4)_2^-$ ,  $Al^{3+}$ ) at  $pH < 3.3$ .  
516 Above this pH limit, however, a limited amount of Al-Fe co-precipitation would still be  
517 possible through variable adsorption of Fe(II) ions onto the formed Al phases.

518 This conceptual figure is consistent with stability diagrams proposed separately for the  
519 Fe(III) and Al systems in previous studies (e.g., Nordstrom, 1982; Bigham et al., 1996;  
520 Bigham and Nordstrom, 2000), and provides a reliable model to account for the behaviour of  
521 Al at low pH. Most notably, it links the mobility of dissolved Al at  $pH < 3.3$  with its interaction  
522 with coexisting Fe(III) phases, and not with discrete Al precipitation in the form of Al phases  
523 such as jurbanite, which has been shown to be irrelevant as a solubility control of Al in acidic  
524 mine waters (Bigham and Nordstrom, 2000; Jones et al., 2011; Sanchez-España et al., 2011).  
525 The incorporation of Al into the Fe(III) phases would be strongly pH-dependent, with  
526 negligible incorporation at very low pH (i.e.,  $pH < 2.0$ ) and increasing Al uptake with  
527 increasing pH (Figs. 5a and 13b).

528 In geochemically varying systems such as these acid pit lakes (pH window of 2.2-4.3  
529 resulting from spatial and temporal variation) the presented geochemical model predicts the  
530 possible coexistence of several mineral sinks of Al, including Al-containing jarosite, Al-  
531 adsorbed schwertmannite, alunite, and hydrobasaluminite. This is valid for environmental  
532 aqueous solutions, as the meta-stability of some of these phases (e.g., schwertmannite with  
533 respect to jarosite and/or goethite, hydrobasaluminite with respect to alunite) will theoretically  
534 tend to significantly reduce this diversity of Al sinks to only one or two stable minerals in the  
535 bottom sediments of the lakes.

#### 536 **4.4. Overall effects of Al-Fe(III) co-precipitation on metal mobility**

537 The combined effects of schwertmannite and jarosite-alunite precipitation on the water  
538 chemistry of the San Temo pit lake are illustrated in Fig. 2. In these vertical profiles of

539 element concentrations, the surface water is variably influenced by seasonal changes (both in  
540 composition and volume) of acid mine drainage inflow seeping from nearby waste piles. The  
541 following observations on metal mobility refer to depths below 10 m depth, where the water  
542 chemistry is mostly controlled by internal hydrogeochemical and geomicrobial processes  
543 occurring in the lake, rather than by external factors.

544 Within a general context of conservative behaviour for most metals (which tend to  
545 increase their concentrations with depth; *see Fig. EA1 in electronic annex*), total iron ( $\text{Fe}_T$ ),  
546 Al, As and Pb are significantly decreased in concentration as a result of their incorporation on  
547 ferric and aluminous phases. The combination of Al-rich jarosite and schwertmannite with  
548 subordinate alunite precipitation promotes a significant decrease of dissolved Al  
549 concentration. Interestingly, the observed removal of aluminum (6-6.7 mM at 25 m depth  
550 compared to 4.7-5.3 mM at a depth of 90 m, representing a 20% net reduction) takes place in a  
551 water column where pH is usually below 3.1 (Fig. 1) and apparently without participation of  
552 any other removal mechanism such as dilution or bioaccumulation. This is noteworthy since  
553 Al is normally considered to be conservative below pH 5.0 in most surface waters.

554 Jarosite and alunite precipitation aids in the removal of total iron (in combination with  
555 schwertmannite; 3.5 mM above the redoxcline vs. 2.5 mM at depth; Fig. 2a). Other elements  
556 which also followed a decreasing vertical trend were arsenic and lead, as well as chromium  
557 and molybdenum to a minor extent (Fig. 2c-d). Schwertmannite normally contains high As  
558 and Pb (up to 3 wt.% As in the Herrerías pit lake; *not shown*). Nevertheless, arsenic has also  
559 been detected in jarosite crystals at trace concentrations by SEM-EDS and STEM-EDS (Fig.  
560 *EA7 in electronic annex*), in addition to variable contents of Pb, Cu and Zn which were very  
561 close to the detection limit and could not be precisely quantified. The uptake of As, Pb, Cu  
562 and Zn by jarosite is supported by ICP-AES analyses of bulk precipitates found at different  
563 depths in ST which were mostly composed of Al-containing jarosite (*Table EA2 in electronic  
564 annex*). The incorporation of As, Pb, Cu and Zn into the crystalline structure of jarosite can be  
565 significant, having important environmental and hydrometallurgical implications (Dutrizac  
566 and Jambor, 2000; Stoffregen et al., 2000; Forray et al., 2014). We could not find conclusive  
567 evidence for the presence of Cr or Mo in jarosite, alunite or schwertmannite, probably due to  
568 the very low concentrations of these two elements in the parent solutions and resulting solids.

## 569 5. CONCLUSIONS

570 The results presented in this study suggest that the behaviour of Al at low pH (<4.0) can  
571 be largely influenced by adsorption on and/or coprecipitation with Fe(III) minerals like  
572 schwertmannite and jarosite. Under conditions typical of AMD and many acid soils and  
573 sediments, alunite formation is feasible and may occur at around pH>3.3, as suggested by  
574 mineralogical observations and geochemical modelling. However, below this pH, and contrary  
575 to the extended assumption, Al is not truly conservative, and in the presence of ferric iron,  
576 both metals may co-precipitate to a substantial extent to form either particles of Al-rich  
577 schwertmannite (containing up to ca. 8 at.% Al with  $[\text{Fe}/(\text{Fe}+\text{Al})]=0.77$ ) and/or crystals of  
578  $\text{H}_3\text{O}^+$ - to  $\text{K}^+$ -jarosite (containing up to ca. 10 at.% Al with  $[\text{Fe}/(\text{Fe}+\text{Al})]=0.54$ ). This Al  
579 incorporation is thought to take place via adsorption on mineral surfaces in schwertmannite  
580 and atomic substitution for  $\text{Fe}^{3+}$  in jarosite. Although macroscopically invisible, the amount of  
581 Al bound to the Fe(III) solid phases may be important enough as to imply a significant Al  
582 removal from the parent solution even at low pH values far below the first hydrolysis point of  
583 this cation ( $\text{pK}_1=4.99$ ). This is especially important in systems with long residence times  
584 favouring metal ion-mineral interaction, such as many pit lakes, flooded mines, tailings or  
585 AMD-affected aquifers.

586 Taken as a whole, the compositional pattern of the analyzed jarosites and alunites  
587 suggests a discontinuous, coupled ( $\text{Al}^{3+}$ - $\text{Fe}^{3+}$ ,  $\text{H}_3\text{O}^+$ - $\text{K}^+$ ) solid solution with an apparent  
588 compositional gap. The lack of intermediate compositions seems to be more closely related to  
589 geochemical restrictions imposed on the precipitation of Fe(III) and Al at equivalent molar  
590 proportions more than with a true immiscibility between the two end-members resulting from  
591 structural limitations in the crystal lattice. It is also envisaged that alunite is unstable with  
592 respect to jarosite at very low pH (<3.3), so that the stability field of this mineral only spans a  
593 narrow pH range (3.3-4.0). Therefore, slight pH shifts taking place in the lakes may lead to  
594 isomorphic replacement of the former mineral by the latter, which may in turn drive the  
595 formation of chemically zoned crystals with jarositic rims around previously formed alunite  
596 cores. The higher degree of  $\text{H}_3\text{O}^+$ - $\text{K}^+$  substitution in jarosites with respect to alunites would be  
597 a natural consequence of the jarosite formation at a substantially lower pH as compared to  
598 alunite.

599 In addition to Al, the Fe(III) solids may also incorporate significant amounts of As and  
600 Pb, so that the co-precipitation of Al and Fe(III) can also exert some control on the fate and  
601 transport of these toxic elements.

602 This work shows that STEM analyses capable of providing chemical information at a  
603 nano-metric resolution can reveal geochemical, mineralogical and water/mineral equilibrium  
604 trends which are often virtually impossible to detect by most other techniques conducted on  
605 bulk samples.

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## 623 **APPENDIX A. ELECTRONIC ANNEX**

624 Supplementary data associated with this article (including analytical data, depth profiles of  
625 metal concentrations, element mapping and additional STEM images of solid phases) can be  
626 found in the online version of this manuscript.

627

## 628 REFERENCES

- 629 Adams F. and Rawajfih Z. (1977) Basaluminite and alunite: a possible cause of sulfate retention by acid soils.  
630 *Soil. Sci. Soc. Am. J.* **41**, 686-692.
- 631 Allison J.D., Brown D.S. and Novo-Gradac J. (1999) MINTEQA2/PRODEAFA2, A geochemical assessment  
632 model for environmental systems: User manual supplement for version 4.0. U.S. EPA, NERL, Athens,  
633 Georgia, 42 pp.
- 634 Anthony J. W. and McLean W. J. (1976) Jurbanite, a new post-mine aluminum sulfate mineral from San Manuel,  
635 Arizona. *Am. Mineral.* **61**, 1-4.
- 636 Alpers C.N., Nordstrom D.K. and Ball J.W. (1989) Solubility of jarosite solid solutions precipitated from acid  
637 mine waters, iron Mountain, California, USA. *Sci. Géol. Bull.* **42-4**, 281-298.
- 638 Alpers C.N., Rye R.R., Nordstrom D.K., White L.D. and King B.S. (1992) Chemical, crystallographic and stable  
639 isotopic properties of alunite and jarosite from acid-hypersaline Australian lakes. *Chemical Geology* **96**, 203-  
640 226.
- 641 Ball J.W. and Nordstrom, D.K. (1991) User's manual for wateq4f, with revised thermodynamic data, base and  
642 test cases for calculating speciation of major, trace, and redox elements in natural waters. USGS Open-File  
643 Report 91-183.
- 644 Baron D. and Palmer C. (1996) Solubility of jarosite at 4-35°C. *Geochimica et Cosmochimica Acta* **60- 2**, 185-  
645 195.
- 646 Bertsch P.M. and Parker B.R. (1996) Aqueous polynuclear aluminum species. In: Sposito G. (Ed.), The  
647 Environmental Chemistry of Aluminum, 2<sup>nd</sup> ed., Lewis Publishers-CRC Press LLC, Boca Raton, Florida, pp  
648 117-168.
- 649 Bigham J.M. and Nordstrom D.K. (2000) Iron and Aluminum Hydroxysulfates from Acid Sulfate Waters. In  
650 Sulfate Minerals: Crystallography, Geochemistry, and Environmental Significance (Alpers, C.N., Jambor,  
651 J.L., Nordstrom, D.K., Eds.). *Reviews in Mineralogy & Geochemistry* **40**, 351-403.
- 652 Blowes D.W., Ptacek C.J. and Jourjovec, J. (2003) Mill tailings: Hydrogeology and Geochemistry. In Jambor  
653 J.L., Blowes D.W. and Ritchie A.I.M. (Eds.) *Environmental Aspects of Mine Wastes*, Mineralogical  
654 Association of Canada, Short Course Series 31, Vancouver, British Columbia, 2003, 95-116.
- 655 Boehrer B., Herzsprung P., Schultze M. and Millero F.J. (2010) Calculating density of water in geochemical lake  
656 stratification models. *Limnol. Oceanogr. Methods* **8**, 567-574.
- 657 Brophy G.P., Scott E.S. and Snellgrove R.A. (1962) Sulfate studies. II. Solid solution between alunite and  
658 jarosite. *Amer. Min.* **47**, 112-126.
- 659 Burger P.V., Papike J.J., Shearer C.K. and Karner J.M. (2009) Jarosite growth zoning as a recorder of fluid  
660 evolution. *Geochim Cosmochim Acta* **73**, 3248-3259.
- 661 Burton E.D., Bush R.T., Johnston S.G., Watling K.M., Hocking R.K., Sullivan L.A., Parker G.K. (2009) Sorption  
662 of Arsenic(V) and Arsenic(III) to Schwertmannite. *Environ. Sci. Technol.* **43-24**, 9202-9207.

- 663 Cornell R.M. and Schwertmann U. (2003) *The Iron Oxides: Structure, Properties, Reactions, Occurrences and*  
664 *Uses* (2<sup>nd</sup> Ed.). Wiley-VCH Verlag, Weinheim, 664 p.
- 665 Davies C.W. (1962) *Ion Association*. London: Butterworths, pp. 37–53.
- 666 Diez-Ercilla M., López-Pamo E. and Sánchez-España J. (2009) The photoreduction of Fe(III) in the acidic mine  
667 pit lake of San Telmo (Huelva, Spain): Field and laboratory work. *Aquatic Geochemistry* **15**, 391-419.
- 668 Diez-Ercilla M., Sánchez-España J., Yusta I., Wendt-Potthoff K. and Koschorreck M. (2014) Formation of  
669 biogenic sulphides in the water column of an acidic pit lake: Biogeochemical controls and effects on trace  
670 metal dynamics. *Biogeochemistry* **121-3**, 519-536.
- 671 Driscoll C.T., Baker J.P., Bisogni J.J. and Schofield C.L. (1984) Aluminum speciation and equilibria in dilute  
672 acidic surface waters of the Adirondack region of New York State. In: *Geological aspects of acid deposition*,  
673 4: 55-75. Bricker OP (Ed.) Ann Arbor Science, Boston.
- 674 Dutrizac J.E. and Jambor J.L. (2000) Jarosites and their application in hydrometallurgy. In *Sulfate Minerals:*  
675 *Crystallography, Geochemistry, and Environmental Significance* (Alpers C.N., Jambor J.L. and Nordstrom  
676 D.K., Eds.). *Reviews Miner. Geochem.* **40**, 405-452.
- 677 Eary L.E. (1999) Geochemical and equilibrium trends in mine pit lakes. *App. Geochemistry* **14**, 963-987.
- 678 Eary L.E. and Castendyk D.N. (2013) Hardrock metal mine pit lakes: Occurrence and geochemical  
679 characteristics. In: Geller, W., Schultze, M., Kleinmann, B., Wolkersdorfer, C. (eds.), *Acidic Pit Lakes: The*  
680 *Legacy of Coal and Metal Surface Mines*. Springer-Verlag, Berlin, Heidelberg, pp. 75–106.
- 681 Falagán C., Sánchez-España J. and Johnson D.B. (2014) New insights into the Biogeochemistry of extremely  
682 acidic environments revealed by a combined cultivation-based and culture-independent study of two stratified  
683 pit lakes. *FEMS Microbiol Ecol* **87-1**, 231-43.
- 684 Forray F.L., Smith A.M.L., Navrotsky A., Wright K., Hudson-Edwards K.A., Dubbin W.E. (2014) Synthesis,  
685 characterization and thermochemistry of synthetic Pb-As, Pb-Cu and Pb-Zn jarosites. *Geochim Cosmochim*  
686 *Acta* **127**, 107-119.
- 687 Friel J.J. (2003) *X-ray and image analysis in electron microscopy*. Princeton Gamma-Tech, 2<sup>nd</sup> Ed., 98 p.
- 688 Garcidueñas R. and Cervantes C. (1996) Microbial interactions with aluminum. *BioMetals* **9**, 311-316.
- 689 Garratt-Reed A.J. and Bell D.C. (2003) *Energy-Dispersive X-Ray Analysis in the Electron Microscope*. BIOS  
690 Scientific Publishers Limited, Oxford, UK, 160 p.
- 691 Härtig C., Brand P. and Bohmhammel K. (1984) Fe-Al isomorphie und strukturwasser in kristallen von Jarosite-  
692 Alunit-Typ. *Z. Anorg. Allg. Chem.* **508**, 159-164.
- 693 Hendershot W.H., Courchesne F. and Jeffries D.S. (1996) Aluminum geochemistry at the catchment scale in  
694 watersheds influenced by acidic precipitation. In: Sposito G. (Ed.), *The Environmental Chemistry of*  
695 *Aluminum*, 2<sup>nd</sup> ed., Lewis Publishers-CRC Press LLC, Boca Raton, Florida, pp 419-449.

- 696 Jamieson H., Robinson C., Alpers C., Nordstrom D.K., Poustovetov A., Lowers H.A. (2005) The composition of  
697 coexisting jarosite-group minerals and water from the Richmond mine, Iron Mountain, California. *The*  
698 *Canadian Mineralogist* **43**, 1225-1242.
- 699 Jones A.M., Collins R.N. and Waite T.D. (2011) Mineral species control of aluminum solubility in sulfate-rich  
700 acidic waters. *Geochim Cosmochim Acta* **75**, 965-977.
- 701 Karathanasis A.D., Evangelou V.P. and Thompson Y.L. (1988) Aluminum and iron equilibria in soil solutions  
702 and surface waters of acid mine watersheds. *J. Environ. Qual.* **17-4**, 534-543.
- 703 Keith W.J., Calk L. and Ashley R.P. (1979) Crystals of coexisting alunite and jarosite, Goldfield, Nevada.  
704 *Shorter Contrib. Miner. Petr.* **1124-C**, C1-C5, Geological Survey Professional Paper.
- 705 Lindsay W.L. and Walthall P.M. (1996) The solubility of aluminum in soils. In: Sposito G. (Ed.), *The*  
706 *Environmental Chemistry of Aluminum*, 2<sup>nd</sup> ed., Lewis Publishers-CRC Press LLC, Boca Raton, Florida, pp  
707 333-361.
- 708 Mosley L.M., Shand P., Self P., and Fitzpatrick R. (2014a) The geochemistry during management of lake  
709 acidification caused by the rewetting of sulfuric (pH<4) acid sulfate soils. *Applied Geochemistry* **41**, 49-56.
- 710 Mosley L.M., Palmer D., Leyden E., Fitzpatrick R., and Shand P. (2014b) Changes in acidity and metal  
711 geochemistry in soils, groundwater, drain and river water in the Lower Murray River after a severe drought.  
712 *Science of the Total Environment* **485-486**, 281-291.
- 713 Newbury D.E. and Ritchie N.W.M. (2013a) Elemental mapping of microstructures by scanning electron  
714 microscopy-energy dispersive X-ray spectrometry (SEM-EDS): extraordinary advances with the silicon drift  
715 detector (SDD). *J. Anal. At. Spectrom.* **28**, 973.
- 716 Newbury D.E. and Ritchie N.W.M. (2013b) Is Scanning Electron Microscopy/Energy Dispersive X-ray  
717 Spectrometry (SEM/EDS) Quantitative?. *Scanning Vol.* **35-3**, 141-168.
- 718 Nordstrom D.K. (1982) The effect of sulfate on aluminum concentrations in natural waters: some stability  
719 relations in the system  $\text{Al}_2\text{O}_3\text{-SO}_3\text{-H}_2\text{O}$  at 298 K. *Geochim. Cosmochim. Acta* **46**, 681-692.
- 720 Nordstrom D.K. and Alpers C.N. (1999) Geochemistry of acid mine waters. In: Plumlee, G.S., and Logsdon,  
721 M.J. (eds.), *The Environmental Geochemistry of Mineral Deposits, Part A. Processes, Techniques, and Health*  
722 *Issues: Society of Economic Geologists, Rev. Econ. Geology* **6A**, 133-156.
- 723 Nordstrom D.K. and Ball J.W. (1986) The geochemical behavior of aluminum in acidified surface waters.  
724 *Science* **232**, 54-56.
- 725 Papike J.J., Karner J.M., Spilde M.N. and Shearer C.K. (2006) Terrestrial analogs of martian sulfates: Major and  
726 minor element systematics of alunite-jarosite from Goldfield, Nevada. *American Mineralogist* **91**, 1197-1200.
- 727 Parkhurst D.L. and Appelo C.A.J. (2013) Description of input and examples for PHREEQC version 3-A  
728 computer program for speciation, batch-reactions, one-dimensional transport, and inverse geochemical  
729 calculations. In: *Groundwater Book 6, Modeling Techniques*. U.S. Geological Survey, Denver, Colorado, p.  
730 497.



- 731 Rancourt D.G. and Ping, J. Y. (1991) Voigt-based methods for arbitrary shape static hyperfine parameter  
732 distributions in Mössbauer spectroscopy. *Nucl. Instrum. Meth. B.* **58-1**, 85-97.
- 733 Regenspurg S. (2002) Characterisation of Schwertmannite - Geochemical Interactions with Arsenate and  
734 Chromate and Significance in Sediments of Lignite Opencast Lakes. PhD Dissertation, University of  
735 Bayreuth, 125 p.
- 736 Regenspurg S. and Peiffer S. (2005) Arsenate and chromate incorporation in schwertmannite. *Applied*  
737 *Geochemistry* **20**, 1226-1239.
- 738 Ripmeester J.A., Ratcliffe C.I., Dutrizac J.E. and Jambor J.L. (1986) Hydronium ion in the alunite-jarosite group.  
739 *Can. Mineral.* **24**, 435-447.
- 740 Sánchez-España J. (2007) The behavior of iron and aluminum in acid mine drainage: Speciation, Mineralogy,  
741 and Environmental Significance. In T.M. Letcher (ed.), *Thermodynamics, Solubility and Environmental*  
742 *Issues*, Elsevier B.V., The Netherlands, pp. 137-149.
- 743 Sánchez-España J., López Pamo E., Santofimia E., Aduvire O., Reyes J. and Baretino D. (2005) Acid Mine  
744 Drainage in the Iberian Pyrite Belt (Odiel river watershed, Huelva, SW Spain): Geochemistry, Mineralogy  
745 and Environmental Implications. *Applied Geochemistry* **20-7**, 1320-1356.
- 746 Sánchez-España J., López-Pamo E., Santofimia E., Reyes J. and Martín Rubí J.A. (2006) The removal of  
747 dissolved metals by hydroxysulfate precipitates during oxidation and neutralization of acid mine waters,  
748 Iberian Pyrite Belt. *Aquat Geochem* **12**, 269-298.
- 749 Sánchez-España J., López-Pamo E., Diez M. and Santofimia E. (2009) Physico-chemical gradients and  
750 meromictic stratification in Cueva de la Mora and other acidic pit lakes of the Iberian Pyrite Belt. *Mine Water*  
751 *Environ* **28**, 15-29.
- 752 Sánchez-España J., Yusta I. and Diez M. (2011) Schwertmannite and hydrobasaluminite: A re-evaluation of their  
753 solubility and control on the iron and aluminum concentration in acidic pit lakes. *Applied Geochemistry* **26**,  
754 1752-1774.
- 755 Sánchez-España J., Yusta I. and López G. (2012) Schwertmannite to jarosite conversion in the water column of  
756 an acidic mine pit lake. *Mineralogical Magazine* **76-7**, 1233-1256.
- 757 Sánchez-España J., Diez M. and Santofimia E. (2013) Mine pit lakes of the Iberian Pyrite Belt: Some basic  
758 limnological, hydrogeochemical and microbiological considerations. In: Geller, W., Schultze, M., Kleinmann,  
759 B., Wolkersdorfer, C. (eds.), *Acidic Pit Lakes: The Legacy of Coal and Metal Surface Mines*. Springer-  
760 Verlag, Berlin, Heidelberg, pp. 315-342.
- 761 Sánchez-España J., Boehrer B. and Yusta I. (2014) Extreme carbon dioxide concentration in acidic pit lakes  
762 provoked by water/rock interaction. *Environ. Sci. Technol.* **48-8**, 4273-4281.
- 763 Singh S.S. (1982) The formation and coexistence of gibbsite, boehmite, alumina and alunite at room temperature.  
764 *Can. J. Soil Sci.* **62**, 327-332.

- 765 Stoffregen R.E., Alpers C.N. and Jambor J.L. (2000) Alunite-Jarosite crystallography, thermodynamics, and  
766 geochronology. In Sulfate Minerals: Crystallography, Geochemistry, and Environmental Significance (Alpers,  
767 C.N., Jambor, J.L., Nordstrom, D.K., Eds.). *Reviews Miner Geochem*, **40**, 453-479.
- 768 Van Breemen N. (1973) Dissolved aluminum in acid sulfate soils and in acid mine waters. *Soil Sci Soc Amer*  
769 *Proc* **37**, 694-697.
- 770 Wendt-Pothoff K., Koschorreck M., Diez M. and Sánchez-España J. (2012) High microbial activity in a  
771 nutrient-rich, acidic mine pit lake. *Limnologica* **42-3**, 175-188.
- 772 Zhu C. and Burden D.S. (2001) Mineralogical compositions of aquifer matrix as necessary initial conditions in  
773 reactive contaminant transport models. *J Contam Hydrology* **51**, 145-161.
- 774

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## TABLE CAPTIONS

776

777 **Table 1.** Chemical composition of San Telmo acidic mine pit lake at several depths, as  
778 measured in July 2010.

779 **Table 2.** Chemical composition of schwertmannite particles formed at different depths in  
780 various acid pit lakes of the IPB (ST, San Telmo, HER, Herrerías, CM, Cueva de la Mora),  
781 as obtained by Energy Dispersion Spectroscopy (EDS) coupled to a Titan3 (FEI)  
782 transmission electron microscope (TEM).

783 **Table 3.** Chemical composition of selected jarosite and alunite crystals formed at different  
784 depths in the San Telmo acid pit lake (ST), as obtained by Energy Dispersion Spectroscopy  
785 (EDS) coupled to a Titan (FEI) transmission electron microscope (TEM). Analyses for  
786 crystals from Cueva de la Mora (CM) are also included. Data in plain text denote whole  
787 crystal analyses, while data in italics represent partial analyses of either cores or rims of  
788 chemically zoned crystals (see Figures 7-8).

789

790

## FIGURE CAPTIONS

791

792 **Figure 1.** Approximate geographical situation (a), field and satellite images (b-c), Oxidation-  
793 reduction potential (ORP) (d) and pH (e) of the San Telmo pit lake, Huelva (SW Spain)  
794 (modified from Sánchez-España et al., 2012). Satellite image taken from Google Earth.

795 **Figure 2.** Vertical profiles of element concentration in the San Telmo acidic pit lake (as  
796 measured in July 2010 –*open symbols*; taken from Sánchez-España et al., 2012– and April  
797 2011 –*black symbols*–): (a) Fe and Al; (b) K and Na, (c) As and Cr, and (d) Mo and Pb. All  
798 element concentrations given in molar proportion.

799 **Figure 3.** Scanning electron microscope (SEM; a-b) and transmission electron microscope  
800 (TEM; c-d) photomicrographs of jarosite and/or alunite crystals formed at different depths  
801 in the water column of the San Telmo (a-b) and Cueva de la Mora (c-d) acid pit lakes. The  
802 concentrations of major elements in the analysed crystals are shown in a-b: (a) Al-rich  
803 jarosite crystal growing in schwertmannite groundmass formed in deep anoxic waters at  
804 100 m depth (note the holes produced by the electronic beam during the EDS analyses); (b)  
805 Isolated crystal of Fe-rich alunite formed at 100 m depth. (c-d) TEM images of

806 concentrically zoned jarosite crystals. Scale bar is 100 nm in (c) and 80 nm in (d).  
807 *Abbreviations: Sch, schwertmannite; Jar, jarosite; Alu, alunite.*

808 **Figure 4.** Chemical mapping of an aggregate of schwertmannite nanoparticles formed at 40 m  
809 depth in the San Telmo acid pit lake, as obtained by Energy Dispersion Spectroscopy  
810 (EDS) with a Transmission Electron Microscope under Scanning mode (STEM): **(a)** High  
811 Angle Annular Dark Field (HAADF) image without any element filter; **(b)** Mapping of S  
812 element distribution; **(c)** Mapping of Fe element distribution; **(d)** Mapping of Al element  
813 distribution.

814 **Figure 5. (a)** Plot of the Fe/[Fe+Al] ratio vs. pH for the analysed schwertmannite particles  
815 from the San Telmo acid pit lake; the error bars are shown for every data point; the inset  
816 shows enlargement of the Y axis to better display the analyses within the range  
817 Fe/[Fe+Al]=0.75-1.00. **(b)** Chemical model of aluminium ionic species distribution for the  
818 pH range 0-8 (computed with PHREEQC for  $a_{\text{SO}_4}=0.1$  M and  $a_{\text{Al}}=0.03$  M; taken from  
819 Sanchez-España, 2007, with permission from Springer). The shaded area represents the pH  
820 window considered in this study. The approximate pH intervals for the observed Fe(III)-Al  
821 coprecipitation and Al adsorption are also indicated.

822 **Figure 6.** Aluminum and iron distributions within a schwertmannite nanoparticle formed at  
823 100 m depth in the San Telmo acid pit lake, as obtained by STEM-EDS: **(a)** High Angle  
824 Annular Dark Field (HAADF) image without any element filter; **(b)** Mapping of Fe and Al  
825 element distribution along with obtained compositions for selected core and edge (needles)  
826 areas (in *at.%*; given as HER65 Particle 6 in Table 2); **(c)** Transversal composition (*line*  
827 *scanning*) across the schwertmannite particle, along with the obtained Al and Fe relative  
828 concentration **(d)**. An outer shell of adsorbed Al (likely composed of  $\text{AlSO}_4^-$  and/or  $\text{AlSO}_4^+$   
829 complexes) is clearly visible around the Fe-containing schwertmannite needles.

830 **Figure 7. (a)** Chemical mapping showing the distribution of aluminium and iron in alunite  
831 (Alu) and jarosite (Jar) crystals formed at 100 m depth in the San Telmo acid pit lake, as  
832 obtained by STEM-EDS; note the different composition of the crystal on the left (in red,  
833 only composed of Al) with respect to the crystal on the right (showing an evident zonation  
834 with an Al-rich core –in red- surrounded by an Fe-rich rim –in blue-; **(b)** *Line scanning*  
835 across the zoned crystal on the right (indicated by a transversal yellow line in (a)) showing  
836 the relative concentration of Al and Fe throughout the crystal. Based on the chemical  
837 composition of selected areas within this crystal (given as *ST100/8\_core* and *ST100/8\_rim*

838 in Table 3) the core is formed by Fe-containing alunite while the rim corresponds to Al-  
839 containing jarosite.

840 **Figure 8.** Chemical mapping of a zoned jarosite crystal from the San Telmo acid pit lake, as  
841 obtained by STEM-EDS: (a) Al; (b) Fe; (c) Al and Fe; (d) Al and Fe superimposed on the  
842 HAADF image, with indication of the area selected as “core” in Table 3  
843 (ST100N14/6\_core); (e) same as (d), with indication of the area selected as “rim” in Table  
844 3 (ST100N14/6\_rim); (f) same as (d), with indication of the *line scan* (transversal yellow  
845 line) across the zoned crystal; (g) transversal scanning shown in (f) showing the relative  
846 concentration of Al and Fe throughout the crystal.

847 **Figure 9.** Chemical mapping of an homogenous alunite crystal from the San Telmo acid pit  
848 lake, as obtained by STEM-EDS: (a) Al; (b) S; (c) Fe; (d) Al and Fe with indication of a  
849 *line scan* (transversal yellow line) conducted across the crystal; (e) same as (d), with  
850 indication of the area selected for the whole crystal analyses labelled as ST100/8\_wcII in  
851 Table 3; (f) K; (g) Results of the line scan shown in (d) with the relative concentration of  
852 Al and Fe throughout the crystal. The iron-rich material outside the alunite crystal is made  
853 of nanometric particles of schwertmannite adhered to the crystal edges.

854 **Figure 10.** High resolution images obtained by transmission electron microscopy (HRTEM)  
855 for outer areas in selected crystals of jarosite (a) and alunite (b). The chemical composition  
856 of these two crystals is also given in Table 3 (ST100-22/1 and ST100N14/9, respectively).  
857 The main diffraction rings and spots are given in the insets, and the corresponding d-  
858 spacings are also indicated.

859 **Figure 11.** Composition of jarosites and alunites formed in the water column of the San  
860 Telmo, Herrerías and Cueva de la Mora acid pit lakes: (a) Binary plot of Al vs. Fe (given as  
861 wt.%); (b) Classification in a plot of  $[\text{Fe}/(\text{Fe}+\text{Al})]_m$  vs.  $[\text{H}_3\text{O}^+]/([\text{H}_3\text{O}^++\text{K}^+)]_m$ . The  
862 compositions have been obtained by SEM-EDS and/or STEM-EDS. The ideal composition  
863 of the pure end-members ( $\text{H}_3\text{O}^+$ -jarosite, *red star*; alunite, *blue star*) is also indicated.

864 **Figure 12.** Vertical evolution of saturation indices for selected mineral phases across the  
865 water column of the San Telmo acid mine pit lake (modified from Sánchez-España et al.,  
866 2012). Calculations computed with PHREEQC using field data and element concentrations  
867 obtained in two different seasons (June 2006 –a– and April 2011 –b–). Abbreviations:  
868 Schw, schwertmannite; Jar- $\text{H}_3\text{O}$ , Hydronian jarosite; Jar-K, Potassium jarosite; Alu,  
869 alunite; Goet, goethite; Gib, gibbsite.

870 **Figure 13. (a)** Plot of the saturation index for selected mineral phases vs. pH in the San Telmo  
871 acid pit lake (modified from Sánchez-España et al., 2012). The samples and chemical data  
872 used to produce this plot are the same as those used in Fig. 12 plus some additional samples  
873 taken in September 2008 from the same lake (Fig. 1) **(b)** pH-pe plot for the Al-Fe-S-K-O-H  
874 system with proposed stability fields for common mineral sinks of dissolved Al. The  
875 Fe(III) species have been also considered to include the incorporation of Al at pH<3.3. The  
876 symbols (white circles, San Telmo; grey triangles, Cueva de la Mora) indicate the pH- pe  
877 conditions observed in waters of different depths in the studied pit lakes. *Abbreviations:*  
878 Schw, schwertmannite; Jar-H<sub>3</sub>O, Hydronian jarosite; Jar-K, Potassium jarosite; Alu,  
879 alunite; Hybs, hydrobasaluminite; Fsbn, felsobanyaite; Gibb, gibbsite.

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**Table 1.** Chemical composition of the San Telmo acid mine pit lake at several depths, as measured in July 2010.

Depth <i>m</i>	Major cations									Major anions			SiO <sub>2</sub> <i>mg/L</i>	DOC <i>mg/L</i>	DIC <i>mg/L</i>
	Na <i>mg/L</i>	K <i>mg/L</i>	Mg <i>mg/L</i>	Ca <i>mg/L</i>	Fe <i>mg/L</i>	Al <i>mg/L</i>	Mn <i>mg/L</i>	Cu <i>mg/L</i>	Zn <i>mg/L</i>	SO <sub>4</sub> <sup>2-</sup> <i>g/l</i>	Cl <sup>-</sup> <i>mg/L</i>	NO <sub>3</sub> <sup>-</sup> <i>µg/L</i>			
0	17	2.3	472	231	114	145	38	18	76	3.6	16	18	58	1.1	35
10	20	1.5	551	296	183	183	49	24	93	4.5	15	16	66	1.2	48
25	20	1.6	579	302	186	182	50	25	96	4.5	14	17	68	1.0	52
30	23	1.3	625	375	164	180	59	27	107	5.1	14	10	63	0.9	114
60	24	1.2	648	425	156	160	65	28	114	4.8	13	6	61	0.8	110
90	25	1.1	666	425	146	145	65	28	114	4.9	14	6	60	0.6	105

Depth <i>m</i>	Trace elements															
	As <i>µg/L</i>	Ag <i>µg/L</i>	Se <i>µg/L</i>	Be <i>µg/L</i>	Ba <i>µg/L</i>	Cd <i>µg/L</i>	Co <i>µg/L</i>	Cr <i>µg/L</i>	Mo <i>µg/L</i>	Ni <i>µg/L</i>	Pb <i>µg/L</i>	Sb <i>µg/L</i>	Th <i>µg/L</i>	Tl <i>µg/L</i>	U <i>µg/L</i>	V <i>µg/L</i>
0	14	b.d.	51	27	16	179	875	14	4	417	56	b.d.	5	3	19	b.d.
10	70	b.d.	72	35	15	243	1.148	21	5	524	71	b.d.	8	3	25	b.d.
25	67	b.d.	69	31	26	245	1.148	21	3	524	70	b.d.	8	3	25	b.d.
30	27	b.d.	75	34	7	267	1.355	18	2	579	44	b.d.	8	3	27	b.d.
60	20	b.d.	78	35	12	287	1.511	17	2	626	19	b.d.	8	3	29	b.d.
90	21	b.d.	78	35	14	288	1.496	17	1	621	19	b.d.	8	3	30	b.d.

b.d., below detection limit (0.2 µg/L)

**Table 2.** Chemical composition of schwertmannite particles formed at different depths in various acid pit lakes of the IPB (ST, San Telmo, HER, Herrerías, CM, Cueva de la Mora), as obtained by Energy Dispersion Spectroscopy (EDS) coupled to a scanning transmission electron microscope (STEM).

Sample	O at. %	Fe at. %	S at. %	Al at. %	Si at. %	Total at. %	$\frac{\text{Fe}}{[\text{Fe}+\text{Al}]}$	Fe/S	pH
ST100ox	66.68	24.19	7.05	0.98	0.4	99.30	0.96	3.43	2.6
ST100ox	65.33	28.08	5.89	0.11	0.11	99.52	1.00	4.77	2.6
ST100ox	66.48	26.92	5.77	0.22	0.13	99.52	0.99	4.67	2.6
ST100ox	66.63	27.11	6.02	0.00	0.00	99.76	1.00	4.50	2.6
CM18	64.06	26.74	6.61	0.70	0.99	99.10	0.97	4.05	2.6
ST35	66.35	25.28	6.59	1.20	0.95	100.37	0.95	3.84	2.6
ST99	65.55	26.11	6.12	1.21	1.27	100.25	0.96	4.27	2.8
ST3.5	64.20	26.35	8.22	1.02	0.00	99.79	0.96	3.21	3.5
HER50 Particle 1	64.90	25.16	5.82	3.16	0.14	99.18	0.89	4.32	4.0
HER50 Particle 2	67.73	23.94	5.00	2.56	0.08	99.31	0.90	4.79	4.0
HER50 Particle 3	66.34	24.76	5.23	2.72	0.08	99.13	0.90	4.73	4.0
HER50 Particle 4	63.51	26.35	5.93	3.37	0.06	99.22	0.89	4.44	4.0
HER50 Particle 5	66.07	24.84	5.49	2.78	0.04	99.22	0.90	4.52	4.0
<b>Ideal Schwertmannite</b>	<b>66.70</b>	<b>29.60</b>	<b>3.70</b>	<b>0.00</b>	<b>0.00</b>	<b>100.00</b>		<b>8.00</b>	
HER65 Particle 6 (core)	58.25	27.61	4.46	8.19	0.24	98.75	0.77	6.19	4.3
HER65 Particle 7 (core)	54.19	31.10	4.82	8.44	0.51	99.06	0.79	6.45	4.3
HER65 Particle 6 (needles)	64.25	11.67	4.51	17.12	1.24	98.79	0.41	2.59	4.3
HER65 Particle 7 (needles)	62.50	14.41	5.08	15.84	0.14	97.97	0.48	2.84	4.3



**Table 3.** Chemical composition of selected jarosite and alunite crystals formed at different depths in the San Telmo acid pit lake, as obtained by Energy Dispersion Spectroscopy (EDS) coupled to a scanning transmission electron microscope (STEM). Analyses for crystals from Cueva de la Mora (CM) are also included. Data in plain text denote whole crystal analyses, while data in italics represent partial analyses of either cores or rims of chemically zoned crystals (see Figures 7-8).

Sample	O <i>at. %</i>	S <i>at. %</i>	Fe <i>at. %</i>	Al <i>at. %</i>	K <i>at. %</i>	Na <i>at. %</i>	Si <i>at. %</i>	Total <i>at. %</i>	Fe [Fe+Al]	H <sub>3</sub> O <sup>+</sup> [H <sub>3</sub> O <sup>+</sup> +K]
CM36/1	68.28	10.96	16.80	0.38	1.26	1.36	0.24	99.28	0.98	0.75
CM36/2	71.97	10.06	15.14	0.55	0.69	1.06	0.26	99.73	0.96	0.86
ST100-22/1	66.22	12.46	15.45	0.51	4.65	0.34	0.85	100.48	0.97	0.09
ST100-22/2	64.65	11.08	16.99	1.00	2.71	0.16	1.75	98.33	0.94	0.47
ST100-22/3	66.90	11.08	17.39	0.73	2.85	0.15	1.05	100.16	0.96	0.44
ST100-22/4	66.25	9.38	16.25	1.88	2.76	0.21	2.91	99.64	0.90	0.46
ST100-22/5	68.21	11.47	16.35	0.75	2.79	0.28	0.58	100.43	0.96	0.45
ST100-22/6	67.24	11.88	19.19	0.30	0.62	0.45	0.07	99.75	0.98	0.88
ST100/1	70.70	10.50	15.38	0.14	3.17	0.11	0.00	100.00	0.99	0.38
ST100/1.2	68.73	10.24	15.37	0.53	2.87	0.66	1.30	99.70	0.97	0.44
ST100/2	65.34	13.16	15.26	0.32	5.11	0.36	0.44	99.99	0.98	0.00
<i>ST100N14/6_rim</i>	<i>65.38</i>	<i>11.47</i>	<i>19.34</i>	<i>0.55</i>	<i>2.68</i>	<i>0.20</i>	<i>0.25</i>	<i>99.87</i>	<i>0.97</i>	<i>0.47</i>
<i>ST100N14/6_core</i>	<i>63.84</i>	<i>12.52</i>	<i>14.19</i>	<i>5.56</i>	<i>3.24</i>	<i>0.31</i>	<i>0.21</i>	<i>99.87</i>	<i>0.72</i>	<i>0.36</i>
<i>ST100/6_rim</i>	<i>61.15</i>	<i>10.21</i>	<i>11.47</i>	<i>9.82</i>	<i>5.06</i>	<i>0.00</i>	<i>1.65</i>	<i>99.36</i>	<i>0.54</i>	<i>0.01</i>
<i>ST100/8_rim</i>	<i>61.12</i>	<i>11.30</i>	<i>16.90</i>	<i>2.86</i>	<i>4.79</i>	<i>1.24</i>	<i>1.52</i>	<i>99.73</i>	<i>0.86</i>	<i>0.06</i>
ST100N14/7	67.44	9.38	4.66	13.11	3.21	0.76	1.41	99.97	0.26	0.20
ST100N14/9	59.74	10.18	2.68	19.19	6.98	0.58	0.51	99.86	0.12	0.13
<i>ST100/6_core</i>	<i>57.98</i>	<i>10.64</i>	<i>4.17</i>	<i>18.08</i>	<i>7.91</i>	<i>0.22</i>	<i>0.78</i>	<i>99.78</i>	<i>0.19</i>	<i>0.01</i>
<i>ST100/8_core</i>	<i>61.75</i>	<i>11.27</i>	<i>7.25</i>	<i>12.43</i>	<i>6.57</i>	<i>0.24</i>	<i>0.35</i>	<i>99.86</i>	<i>0.37</i>	<i>0.18</i>
<i>ST100/8_core2</i>	<i>61.55</i>	<i>10.92</i>	<i>5.56</i>	<i>14.15</i>	<i>6.89</i>	<i>0.47</i>	<i>0.46</i>	<i>100.00</i>	<i>0.28</i>	<i>0.14</i>
ST100/8_wc	62.63	10.24	7.24	12.47	6.55	0.25	0.36	99.74	0.37	0.18
ST100/8_wcll	63.25	10.35	1.37	17.80	6.19	0.11	0.73	99.80	0.07	0.23
<b>Jarosite (ideal)</b>	70.00	10.00	15.00	0.00	5.00	-	-	100.00	1.00	
<b>H-Jarosite (ideal)</b>	75.00	10.00	15.00	0.00	0.00	-	-	100.00	1.00	
<b>Alunite (ideal)</b>	70.00	10.00	0.00	15.00	5.00	-	-	100.00	0.00	

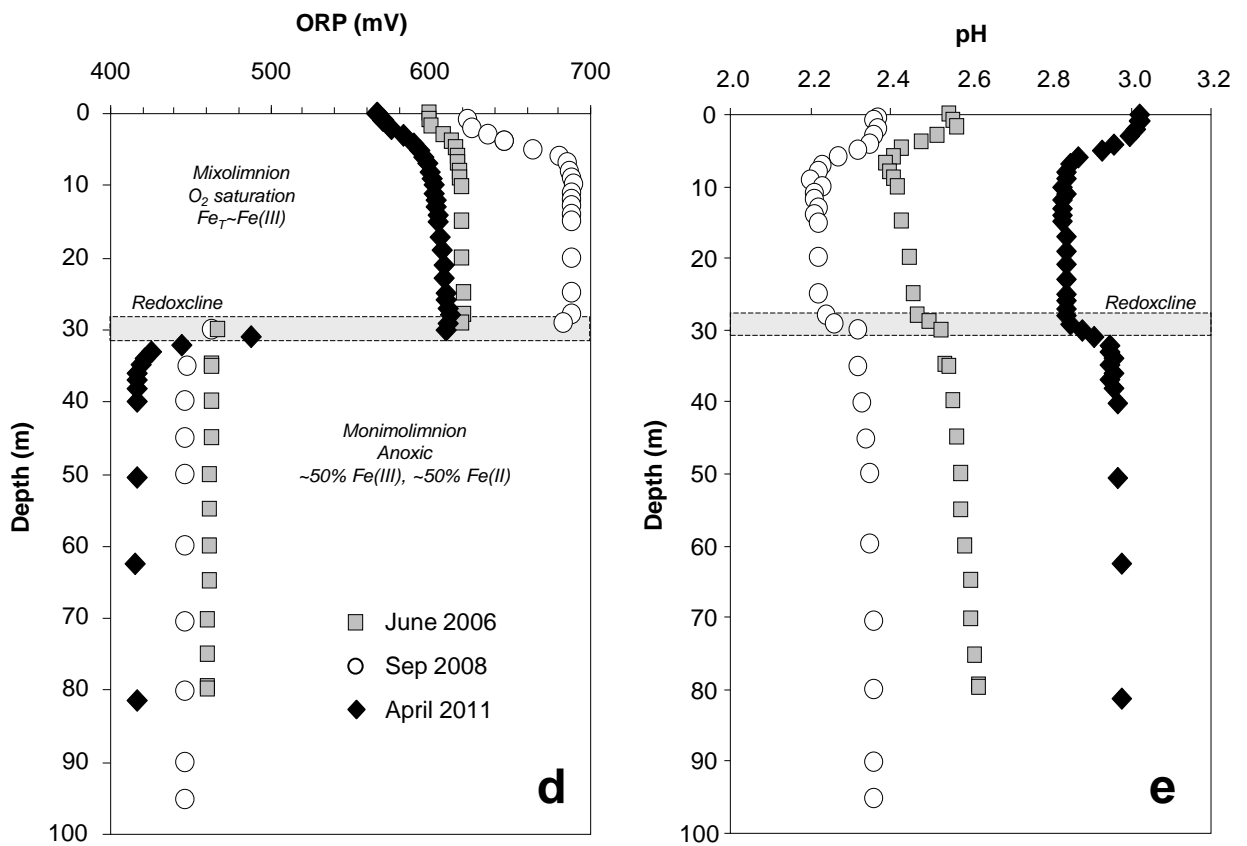
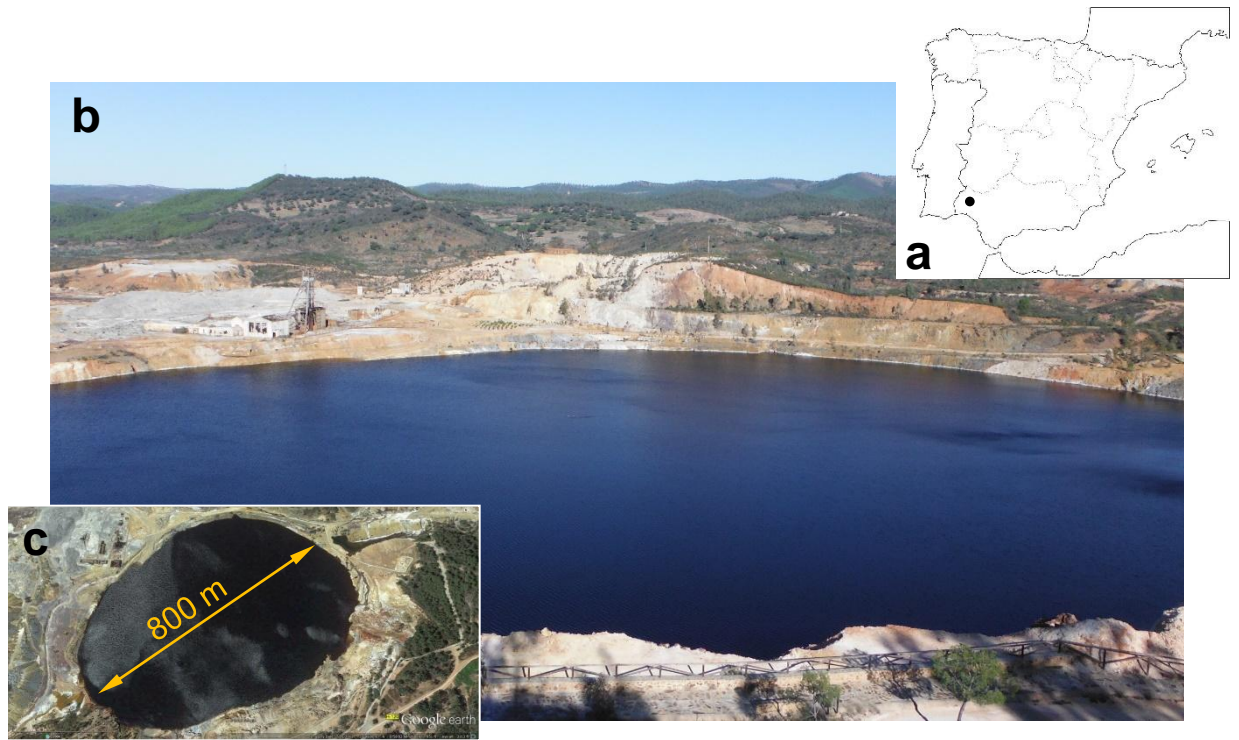


Figure 1

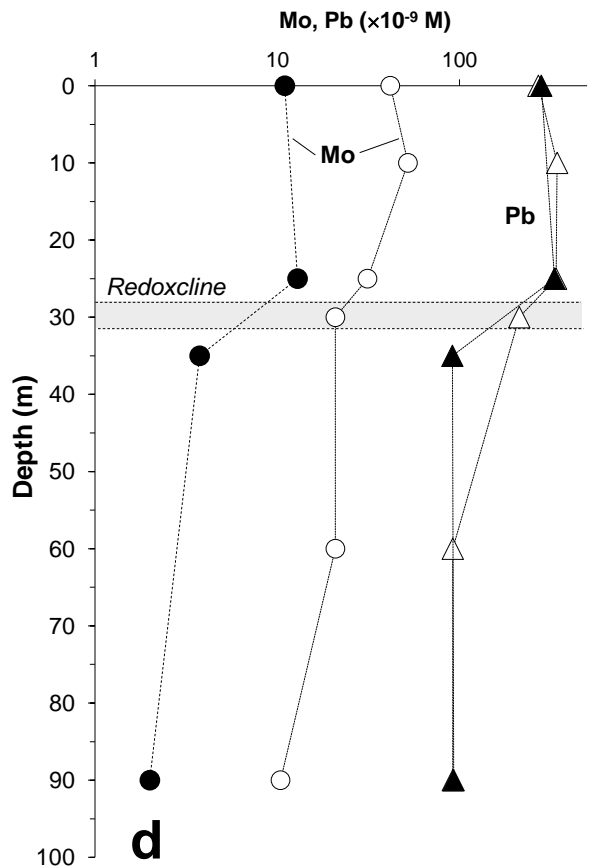
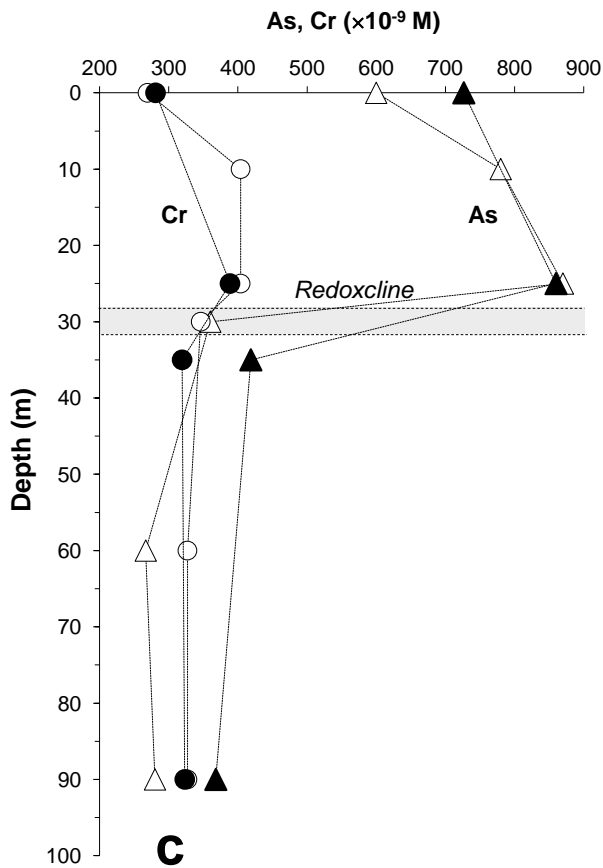
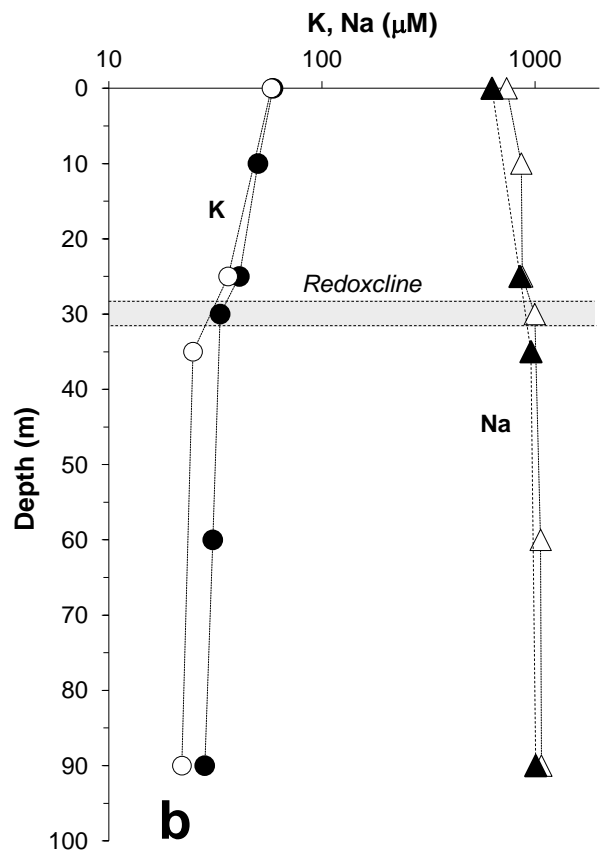
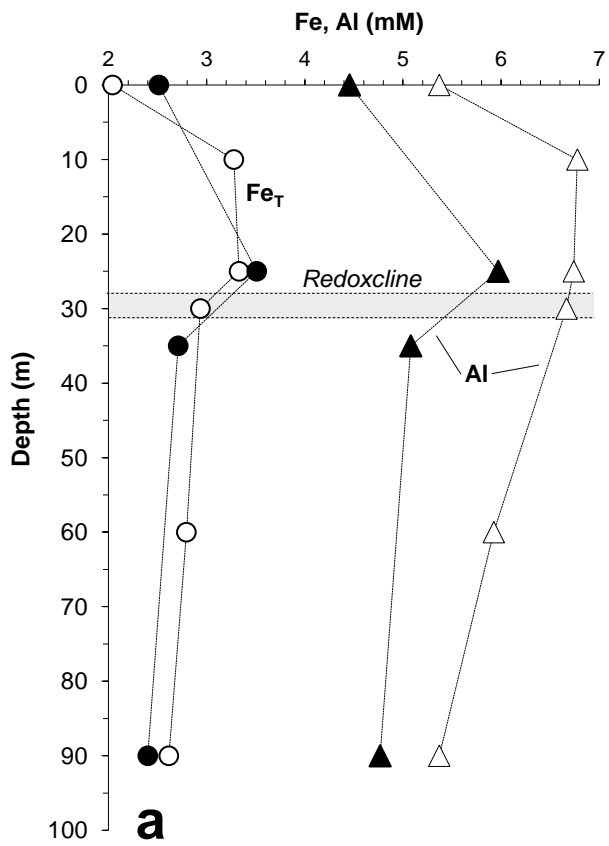


Figure 2

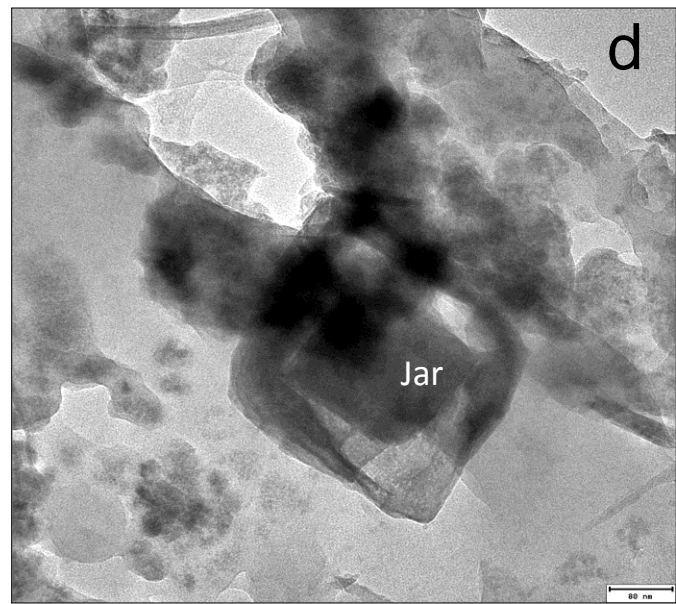
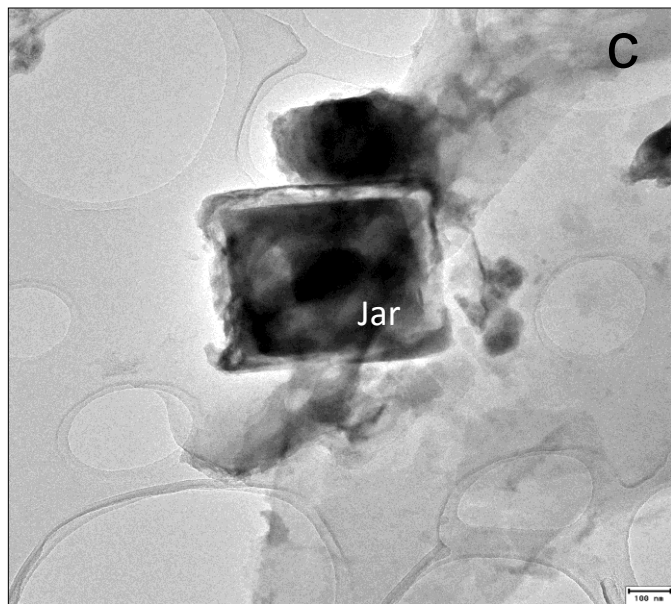
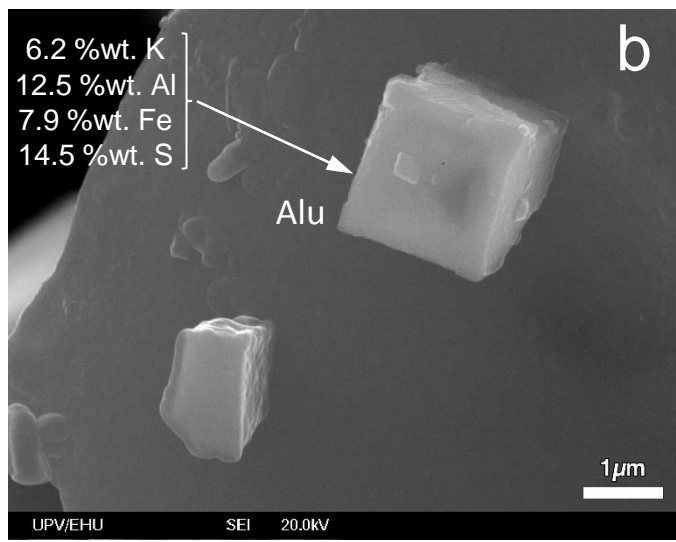
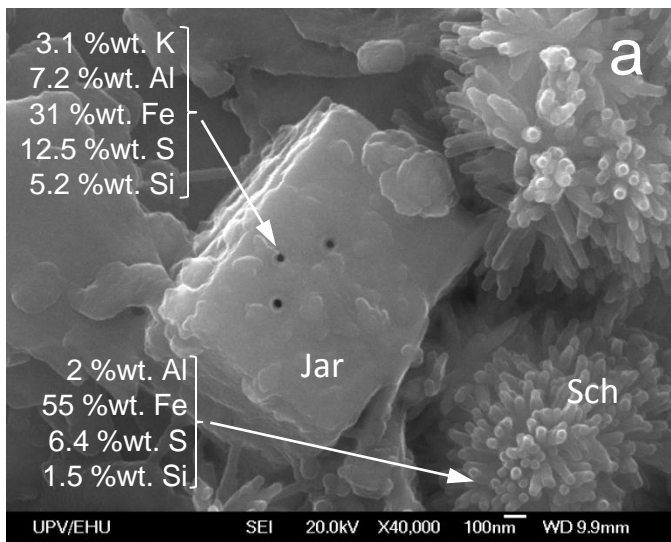


Figure 3

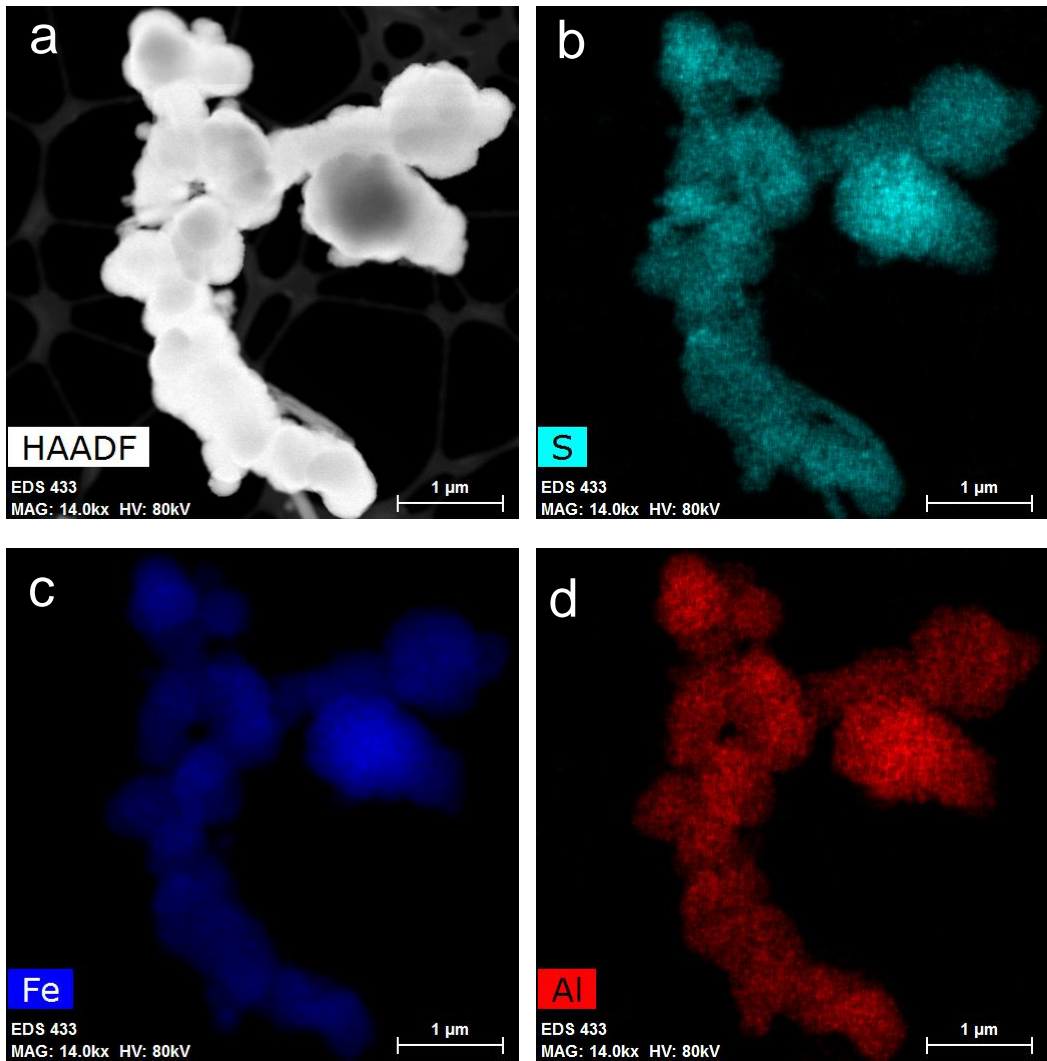
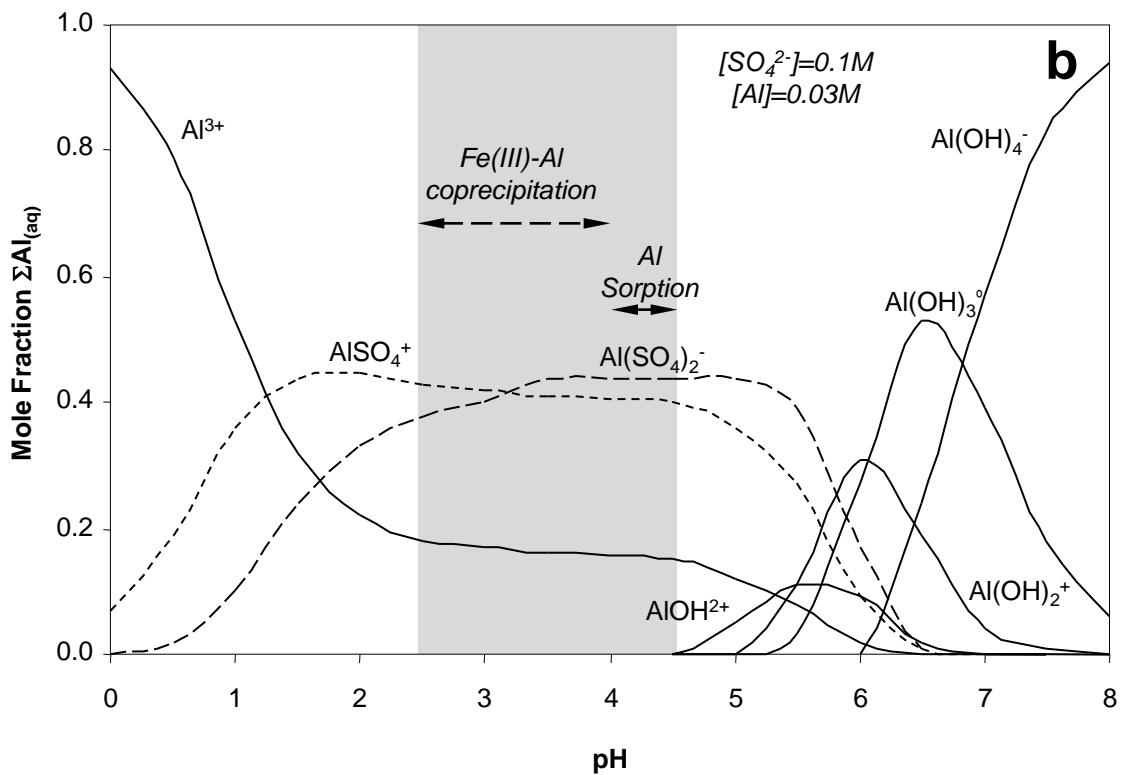
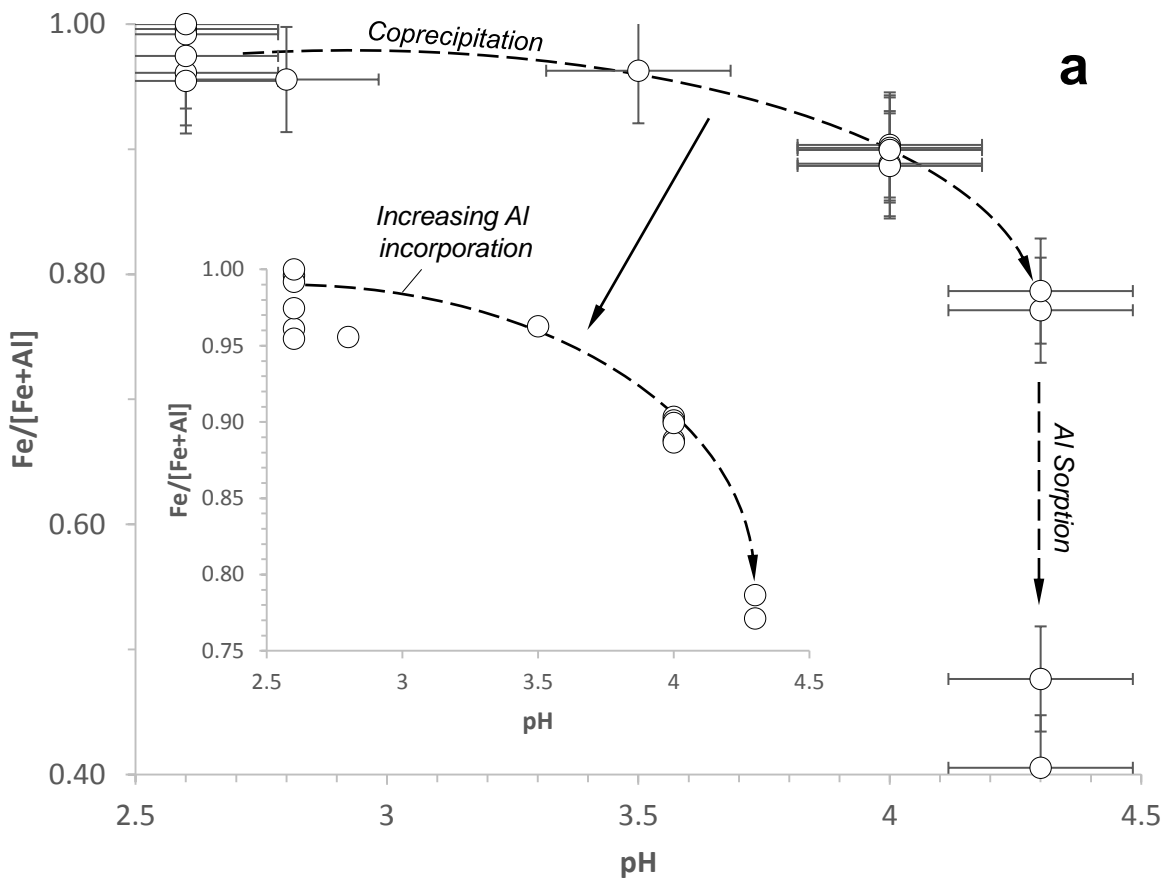
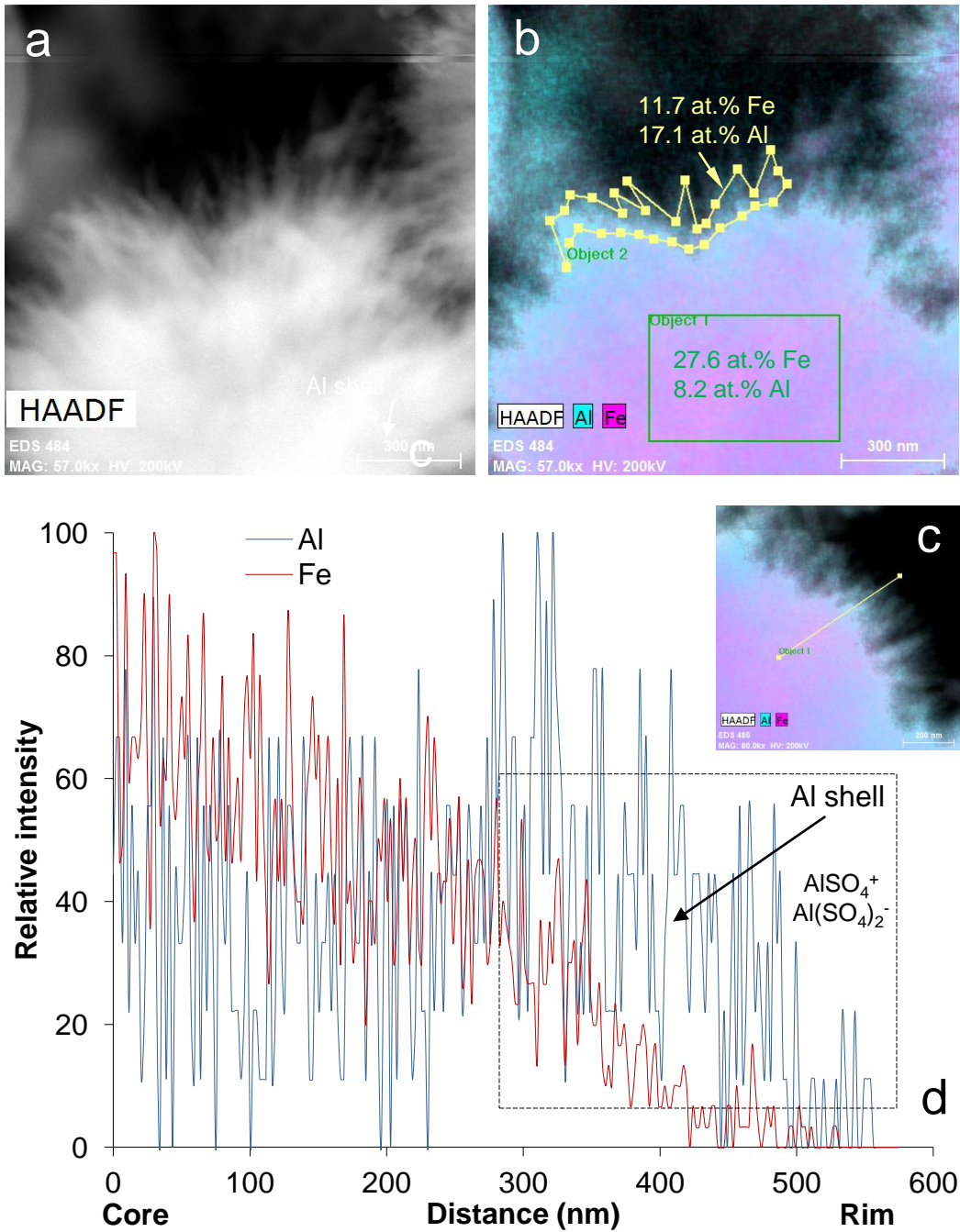


Figure 4



**Figure 5**



**Figure 6**

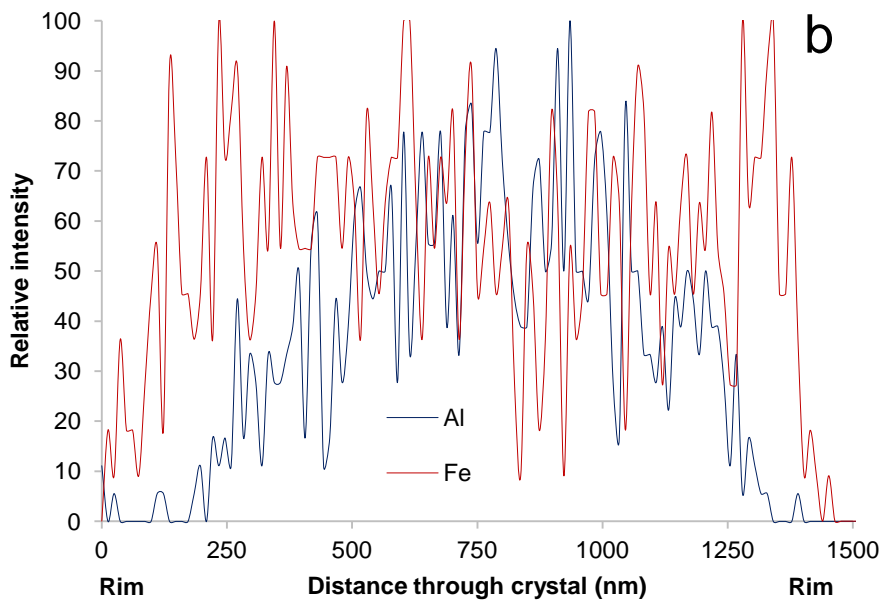
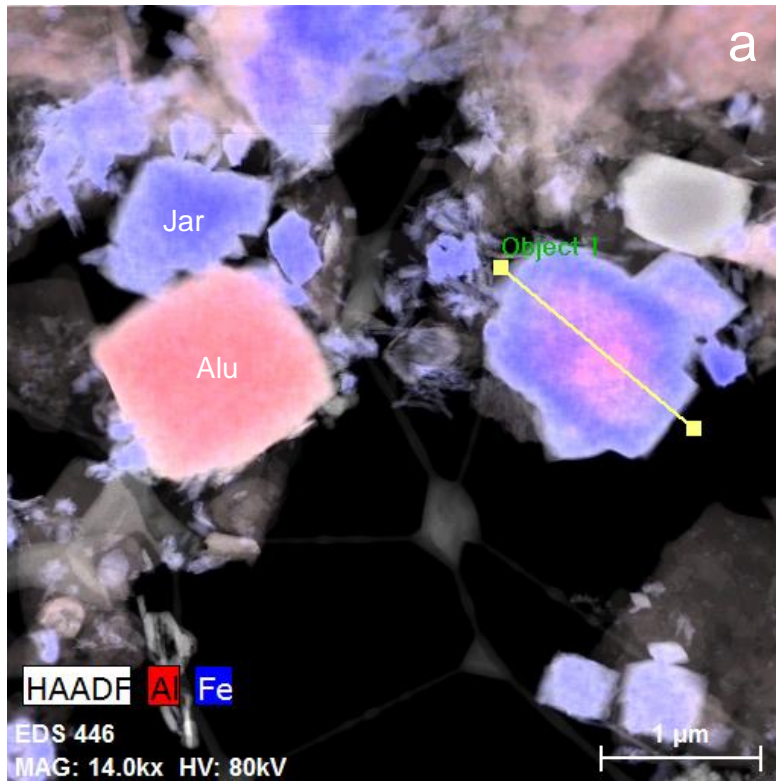


Figure 7



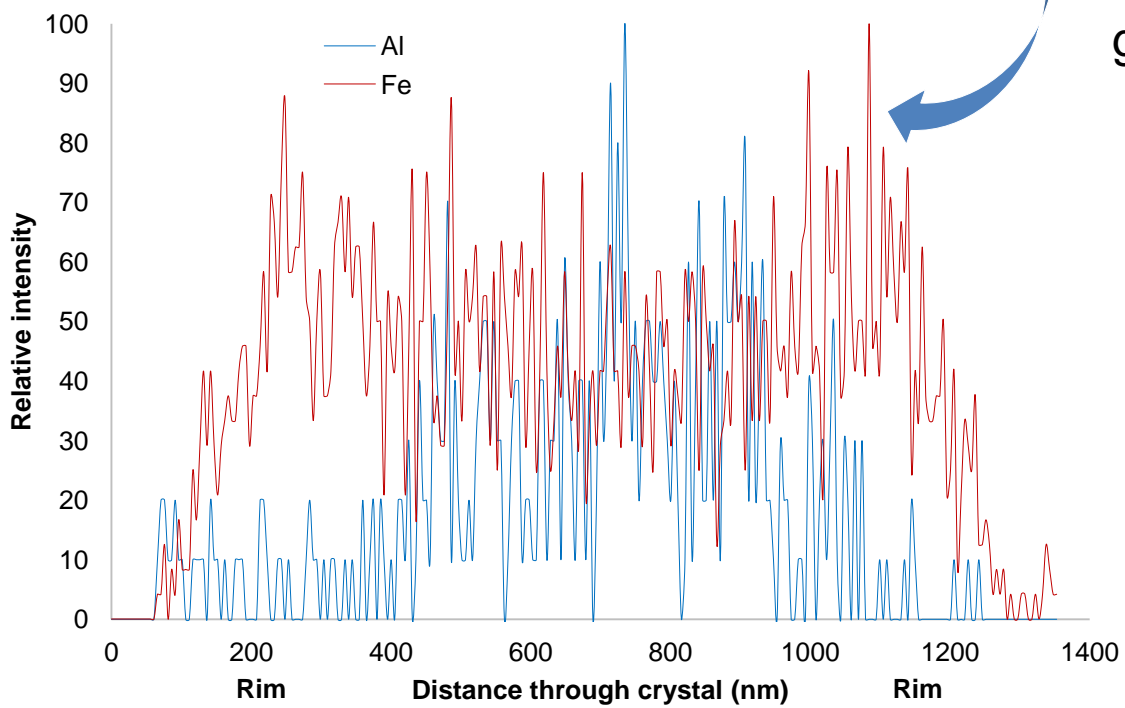
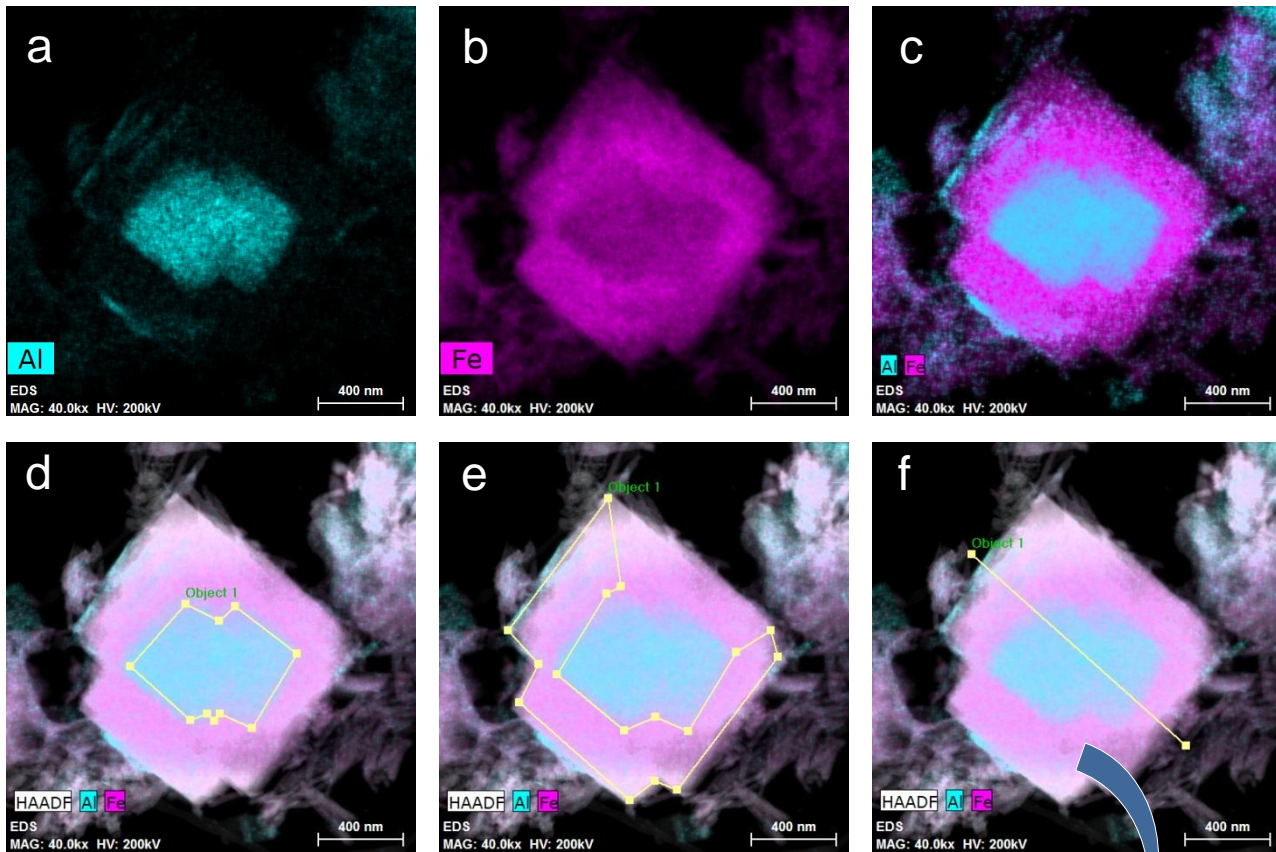
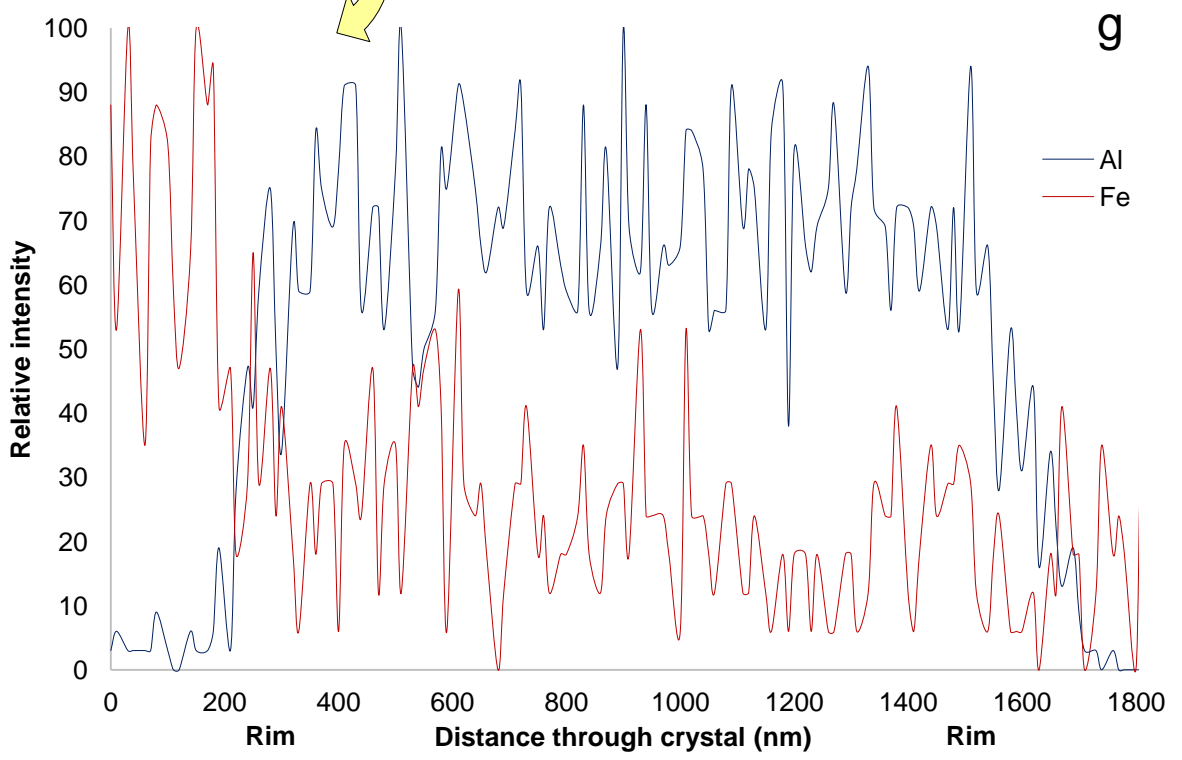
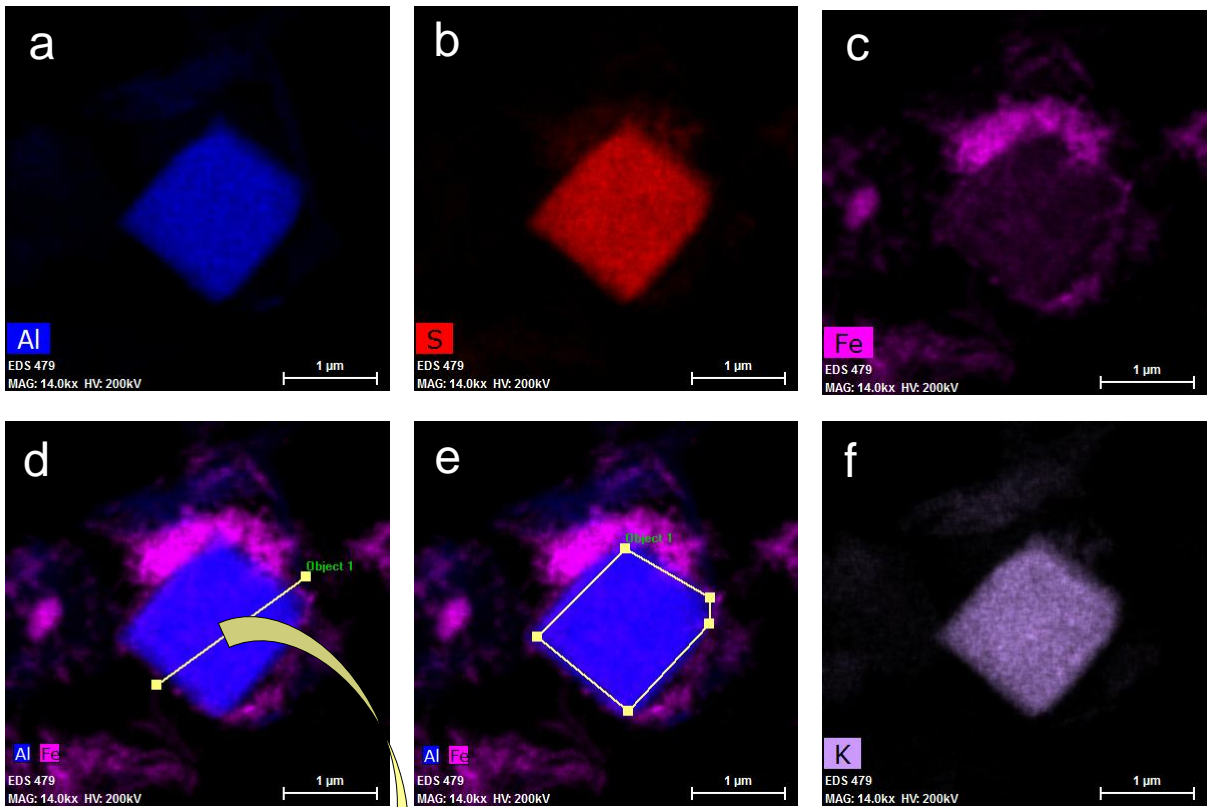
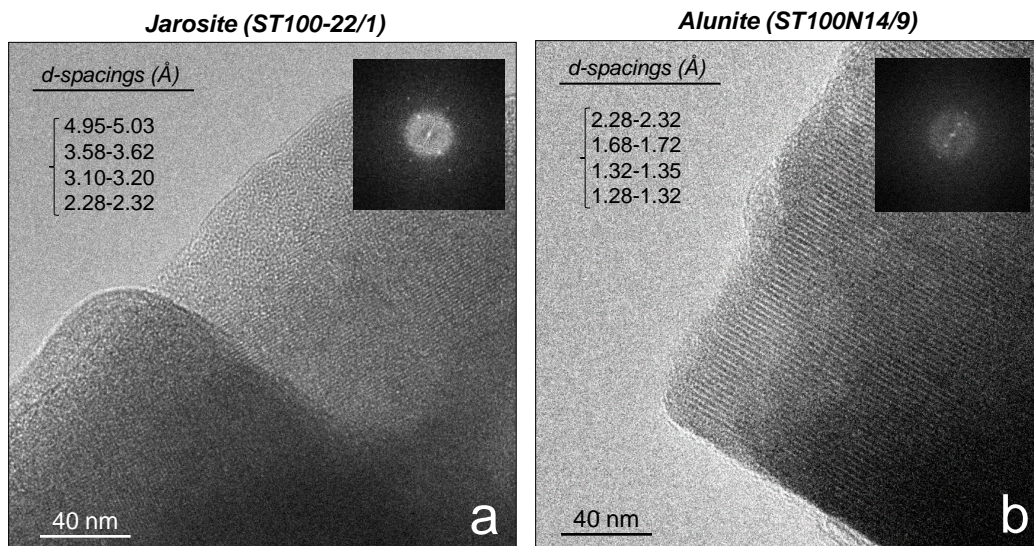


Figure 8



**Figure 9**



**Figure 10**

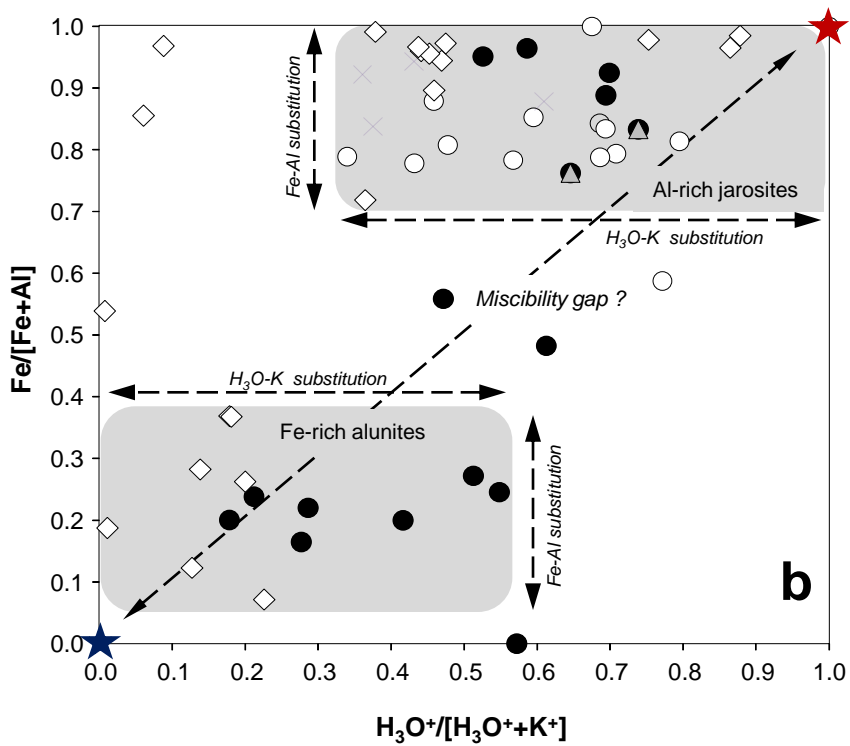
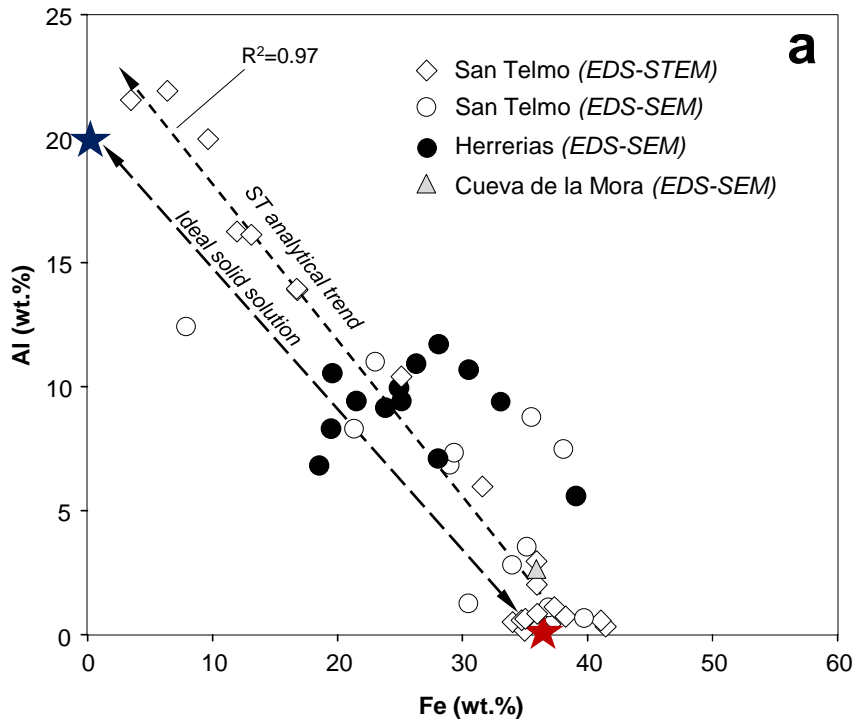
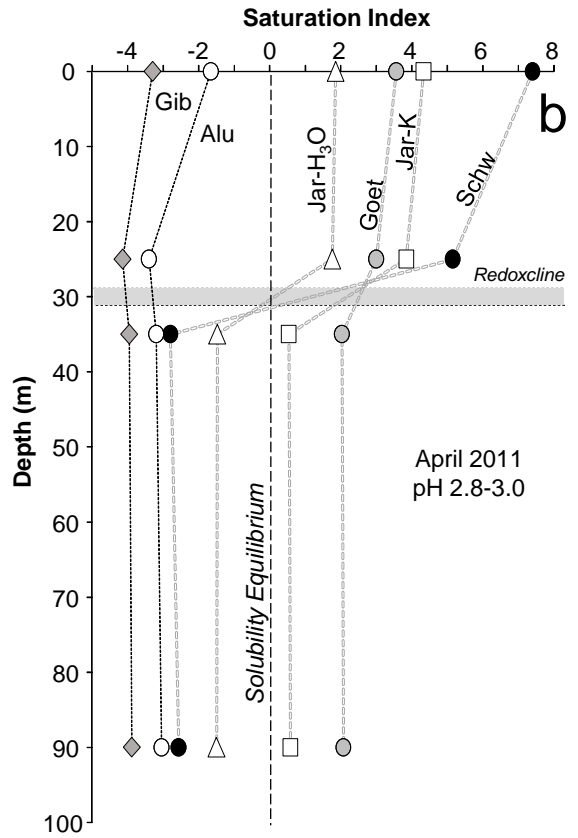
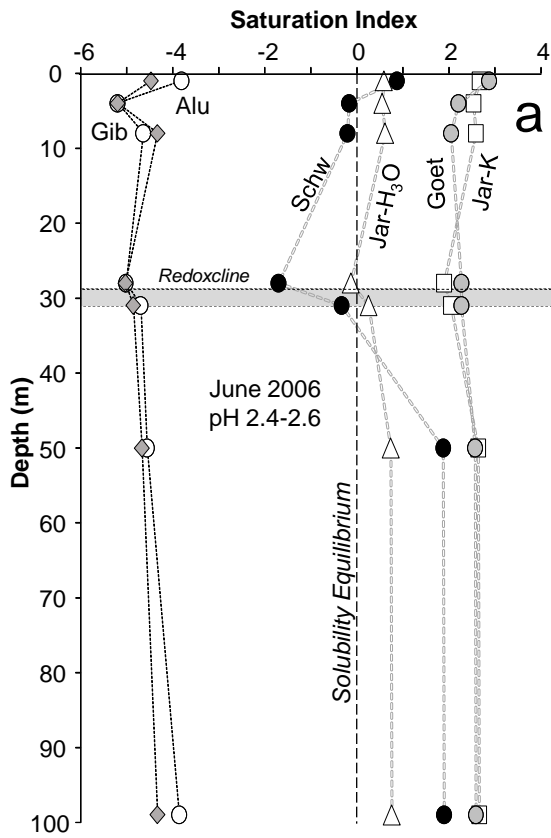
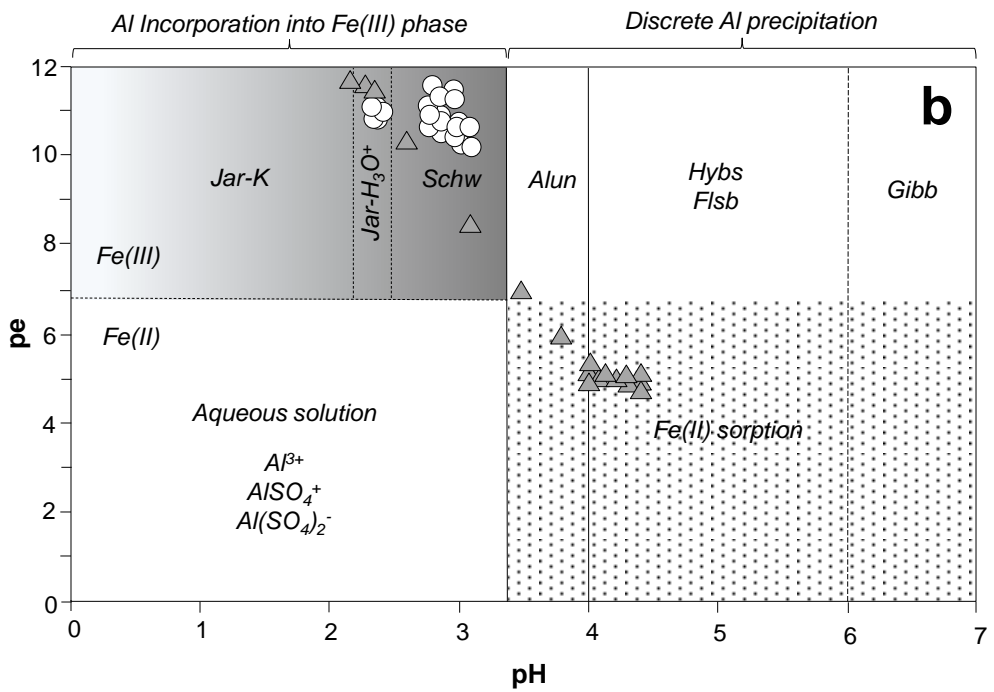
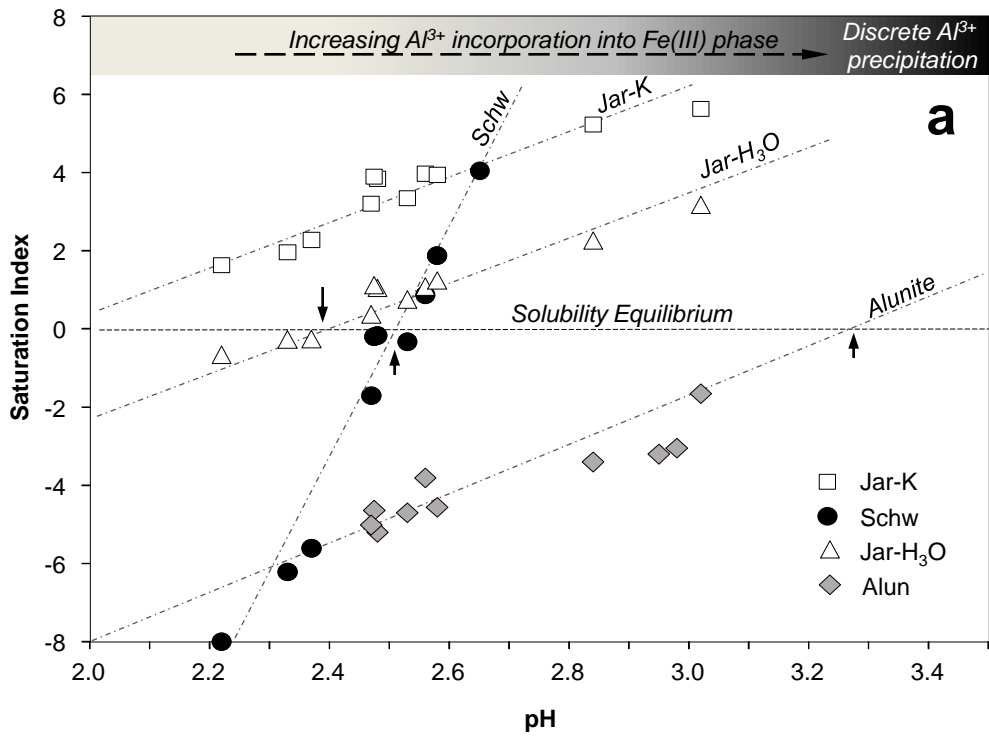


Figure 11



**Figure 12**



**Figure 13**